

1 *Communication*

2 **Ecological drivers of species distributions and niche
3 overlap for three eastern subterranean termite species
4 in the southern Appalachian Mountains, USA**

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8 **Abstract:** In both managed and unmanaged forests, termites are functionally important members of
9 the dead-wood-associated (saproxylic) insect community. However, little is known about regional-
10 scale environmental drivers of geographic distributions of termite species, and how these
11 environmental factors impact co-occurrence among congeneric species. Here we focus on the
12 southern Appalachian Mountains—a well-known center of endemism for forest biota—and use
13 Ecological Niche Modeling (ENM) to examine the distributions of three species of *Reticulitermes*
14 termites (i.e., *R. flavipes*, *R. virginicus*, and *R. mallei*). To overcome deficiencies in public databases,
15 ENMs were underpinned by field-collected high-resolution occurrence records coupled with
16 molecular taxonomic species identification. Spatial overlap among areas of predicted occurrence of
17 each species was mapped, and aspects of niche similarity were quantified. We also identified
18 environmental factors that most strongly contribute to among-species differences in occupancy.
19 Overall, we found that *R. flavipes* and *R. virginicus* showed significant niche divergence, which was
20 primarily driven by dry-season precipitation. Also, all three species were most likely to co-occur in
21 the mid-latitudes of the study area (i.e., northern Alabama and Georgia, eastern Tennessee and
22 western North Carolina), which is an area of considerable topographic complexity. This work
23 provides important baseline information for follow-up studies of local-scale drivers of these species'
24 distributions. It also identifies specific geographic areas where future assessments of the frequency
25 of true syntopy vs. micro-allopatry, and associated interspecific competitive interactions, should be
26 focused.

27 **Keywords:** Appalachian Mountains; biodiversity; ecological niche model; PCR-RFLP; molecular
28 taxonomic identification; niche overlap; *Reticulitermes*; saproxylic; species distribution; termite
29

30 **1. Introduction**

31 **1.1. The southern Appalachian Mountains: a center of endemism for forest biota**

32 The southern Appalachian Mountains, extending latitudinally from northeast Alabama to
33 northwest Virginia, are some of the oldest uplands in North America. These mountains have been
34 exposed and unglaciated for over 100 million years [1]. Steep altitudinal precipitation gradients, a
35 complex heavily dissected topography, and a humid, temperate climate, have shaped southern
36 Appalachian forests into some of the most diverse environments in the eastern United States. [2].
37 While deciduous oak-hickory forests dominate much of the mid-elevation landscape [2], high
38 elevations (above 1400 m) support spruce-fir forests [3], whereas mesic coves support hemlock, and
39 pines are commonly found at xeric low- to mid-elevations [4].

40 The southern Appalachian Mountains are incredibly rich in biodiversity [5]. The region is
41 thought to have served as a major Pleistocene refuge for numerous species. Past climatic cycles have
42 affected distributions of forest biota, resulting in major range shifts or local extinction. Following the
43 Last Glacial Maximum (ca. 21,000 years ago), recolonization is thought to have occurred relatively

44 rapidly, from 7,000-16,000 years ago [6–10]. The southern Appalachian Mountains are a well-known
45 center of endemism for salamanders and other amphibians [11,12]. However, there is increasing
46 evidence of short-range endemism in other groups, including dead wood-associated forest
47 invertebrates (e.g., millipedes [13,14], cockroaches [15,16], and centipedes [17]).

48 *1.2. Subterranean termites: functionally important ecosystem service providers in temperate forests*

49 Dead-wood-dependent (saproxylic) arthropods play critical roles in maintaining healthy,
50 productive forests by contributing to the decomposition of fallen trees and thus driving nutrient
51 cycling that affects organisms at all trophic levels [18–22]. Indeed, rotting logs may be one of the most
52 stable, thermally buffered, above-ground microhabitats that exist in forests, and the decomposition
53 process has successional stages, facilitated by wood-feeding and wood-boring invertebrates [21,23].
54 Termites are some of the first to colonize a rotting log, and through feeding and tunneling activities
55 of the worker caste, the dead-wood substrate is modified by the creation of galleries. Once
56 established, these facilitate colonization by larger wood-feeding invertebrates [24]. Ultimately, the
57 ecosystem engineering activities of termites contribute to enhancing the internal heterogeneity of
58 logs, making them habitable by a diverse array of saproxylic species.

59 Termites in the genus *Reticulitermes* (Blattodea: Rhinotermitidae) are broadly distributed across
60 the eastern United States. Morphological separation of species is notoriously difficult [25],
61 particularly given that only the worker caste can usually be readily sampled. To address this, we
62 developed an efficient molecular assay (i.e., polymerase chain reaction (PCR) amplification of a short
63 region of mitochondrial cytochrome oxidase subunit II gene, followed by screening of restriction-
64 fragment-length polymorphism (RFLP) banding profiles [26]) that can be used to distinguish each of
65 the five eastern United States species. In the southern Appalachians, several *Reticulitermes* species can
66 co-occur locally. However, true syntopy (i.e., two species co-inhabiting the same rotting log) appears
67 to be very rare, and reported instances of fine-scale sampling have been limited.

68 *1.3. Ecological niche models: efficient tools for predicting organismal distributions*

69 Ecological niche models (ENMs) are broadly useful spatially explicit analytical tools that relate
70 species occurrence data with climate variables, such as temperature and precipitation [27]. Once
71 constructed, ENMs generate maps of estimated habitat suitability, and can be used to describe the
72 historical, current, and future climate space for a given species. For example, ENMs have been used
73 to identify areas of high conservation importance, predict climate change effects on geographic
74 ranges of species, as well as determine potential threats of invasive species [28]. These analytical tools
75 are becoming widely used owing to the increasing accessibility of climatic data via public databases
76 [29–31]. An important assumption when using ENMs to predict historical or future species'
77 distributions is niche conservatism (i.e., the stability of ecological niches over time) [28]. However,
78 evidence suggests that niche conservatism is common among closely related species [32–34], and the
79 risks of erroneous inferences are further reduced when focusing only on contemporary climate and
80 occurrence data (i.e., reconstructing present-day ENMs).

81 *1.4. The current state of knowledge about subterranean termite distributions, and goals of this study*

82 There is a general lack of data on the natural distributions of termites in temperate forests, given
83 that most research has focused on damage that termites cause to man-made wooden structures.
84 Accordingly, occurrence records mostly come from urban areas, and they are also of low resolution
85 (e.g., presence/absence in a given county). Notwithstanding these limitations, Maynard et al. [35]
86 recently provided valuable insights into the role of climatic (temperature and precipitation) variables
87 in influencing distributions of termites in the eastern United States. Specifically, those authors
88 performed ENM for two *Reticulitermes* species (*R. flavipes* and *R. virginicus*) and the invasive
89 Formosan subterranean termite, *Coptotermes formosanus*. Furthermore, they synthesized pre-existing
90 knowledge to identify the influence on termite distributions of biotic factors, such as tree species and
91 wood traits, fungal preferences, phenology of predatory ants, and competitive asymmetries among

92 coexisting termite species. While interspecific competition may result in spatial or temporal
93 separation which could lead to niche divergence, to date, very little is known about niche partitioning
94 in subterranean termites and the environmental factors that may lead to niche conservatism or
95 divergence.

96 In the present paper we aimed to generate new insights into regional-scale environmental
97 drivers of geographic distributions of termite species, and how these environmental factors impact
98 co-occurrence among congeneric species. Focusing on the southern Appalachian Mountains and
99 surrounding areas, we performed an ENM-based evaluation of niche divergence among the three
100 most common *Reticulitermes* species in the eastern United States. In addition to identifying niche
101 divergence, if present, we aimed to determine the environmental factors driving niche divergence
102 among species.

103 2. Materials and Methods

104 2.1. Termite sampling, species identification, and Ecological Niche Modeling

105 From 2012 to 2016, we collected *Reticulitermes* termites from 132 sites across the southern
106 Appalachians Mountains and surrounding areas (Table S1; Figure S1). At most sites, termite workers
107 were collected from a single rotting log at an intermediate to late stage of decay. However, at 10 sites,
108 termites were also collected from additional logs within ~30 m of one another (i.e., samples came
109 from a total of 2 logs at 8 sites, 3 logs at 1 site, and 4 logs at 1 site; Table S1). Owing to the close
110 proximity of these clustered logs (i.e., at or near the typical error associated with a handheld GPS
111 unit), the same coordinates were assigned to them, but specimen collections were assigned log-
112 specific identifiers. Molecular taxonomic identifications were based on a single termite per rotting
113 log, using Garrick et al.'s [26] PCR-RFLP assay. Ultimately, we identified 91 non-redundant
114 occurrence points for *R. flavipes*, 30 for *R. virginicus*, and 17 for *R. mallei* (Table S1). ENM was
115 conducted using the 'biomod2' package [36] in R [37] using four modeling algorithms. Present-day
116 distributions were based on mean climatological data spanning 1960–1990. Nineteen bioclimatic
117 variables [29] were obtained from the WorldClim database v.1.4 (<http://www.worldclim.org>), and
118 then factor analysis was used to reduce the number of predictors, and the associated correlation
119 among them (see Supporting File 1 for full details of ENM methods). From the nineteen bioclimatic
120 variables, we generated four environmental factors (see Supporting File 1 for full details of factor
121 analysis): dry-season precipitation, wet-season precipitation, summer temperature, and temperature
122 range.

123 2.2. Niche occupancy, niche identity, and distributional overlap

124 Predicted niche occupancy profiles were generated for each environmental factor following
125 Evans et al. [38], implemented in the 'phyloclim' package [39]. Niche overlap for each environmental
126 factor was summarized using both Schoener's D statistic [40], and the modified Hellinger statistic, I,
127 as proposed by Warren et al. [41]. We also used the D and I statistics to determine pairwise niche
128 equivalency/identity among the three *Reticulitermes* species. The niche equivalency test asks whether
129 the ENMs of two species are more different than expected if they had been drawn from the same
130 distribution. To perform the niche equivalency test, we generated a distribution using 999
131 pseudoreplicate datasets.

132 To assess distributional overlap, we converted habitat suitability maps for each species to binary
133 presence/absence maps and then overlaid these binary maps on one another. To convert continuous
134 occurrence probabilities (i.e., habitat suitability) to binary classifications, we used a threshold value
135 of 0.2, where suitability > 0.2 represented presence, and suitability ≤ 0.2 represented absence. This
136 occurrence probability threshold was chosen based on the True Skill Statistic [42]. We overlaid the
137 three species' binary maps by summing re-coded maps, where absence = 0, but presence was coded
138 depending on species: *R. flavipes* = 4, *R. virginicus* = 2, and *R. mallei* = 1. This way, the sum of binary
139 maps resulted in seven distinct categories: single-species areas (3 categories, with aforementioned

140 scores); areas of two-species overlap (3 categories, scores of either 3, 5, or 6 depending on the identity
141 of the species pair); and areas where all three species overlap (1 category, with a score of 7).

142 *2.3. Environmental factors and niche divergence*

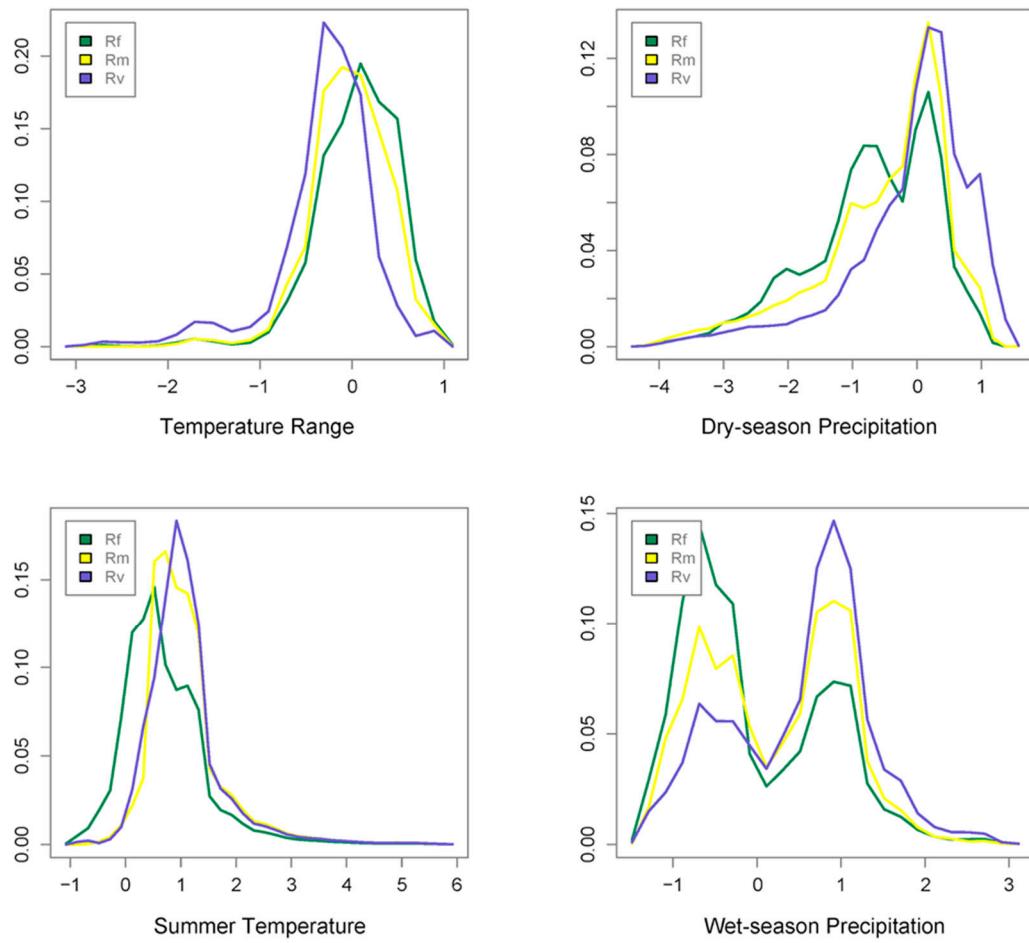
143 In order to determine the sources of variation in the *Reticulitermes* occurrence dataset, we
144 included the effects of spatial structure and environmental factors, and performed variance
145 partitioning using the 'varpart' function in 'vegan' [44]. To account for multiple predictors in the
146 model, we used adjusted R^2 . To determine which (if any) environmental factors have significantly
147 contributed to niche divergence of *Reticulitermes* species, we first removed the effect of spatial
148 structure. We did this by performing distance-based redundancy analysis [43] using the 'capscale'
149 function. To account for spatial structure, we transformed Euclidean geographic distances to a
150 continuous rectangular vector by Principal Coordinates analysis of Neighbor Matrices (PCNM) using
151 the 'pcnm' function in 'vegan'. Only significant PCNM axes were used in partialling out spatial
152 structure. Significance of the environmental and spatial predictors was assessed using multivariate
153 F-statistics with 9999 permutations.

154 **3. Results**

155 *3.1. Niche occupancy, niche identity, and distributional overlap*

156 Predicted niche occupancy profiles for the three *Reticulitermes* species (Figure 1) showed
157 differences in peak values across all four environmental factors. The two temperature factors,
158 summer temperature and temperature range, showed differences in peaks between *R. flavipes* and *R.*
159 *virginicus*, whereas *R. mallei* was intermediate. Similarly, the two precipitation factors, dry-season
160 precipitation and wet-season precipitation, showed more marked differences between *R. flavipes* and
161 *R. virginicus* than for any of the other pairwise species comparisons. Statistics that characterize the
162 extent of niche overlap showed the least amount of niche overlap was between *R. flavipes* and *R.*
163 *virginicus* ($D = 0.582$, $I = 0.843$; Table 1). Furthermore, the niche identity test between these two species
164 showed significant differentiation ($P < 0.001$; Table 1). *R. mallei* was more similar to *R. flavipes* in
165 terms of temperature range ($D = 0.889$) and dry-season precipitation ($D = 0.872$), but showed more
166 overlap with *R. virginicus* for summer temperature ($D = 0.894$) and wet-season precipitation ($D =$
167 0.848). *R. virginicus* showed the least overlap with *R. flavipes*, across all four environmental factors
168 (Table S2).

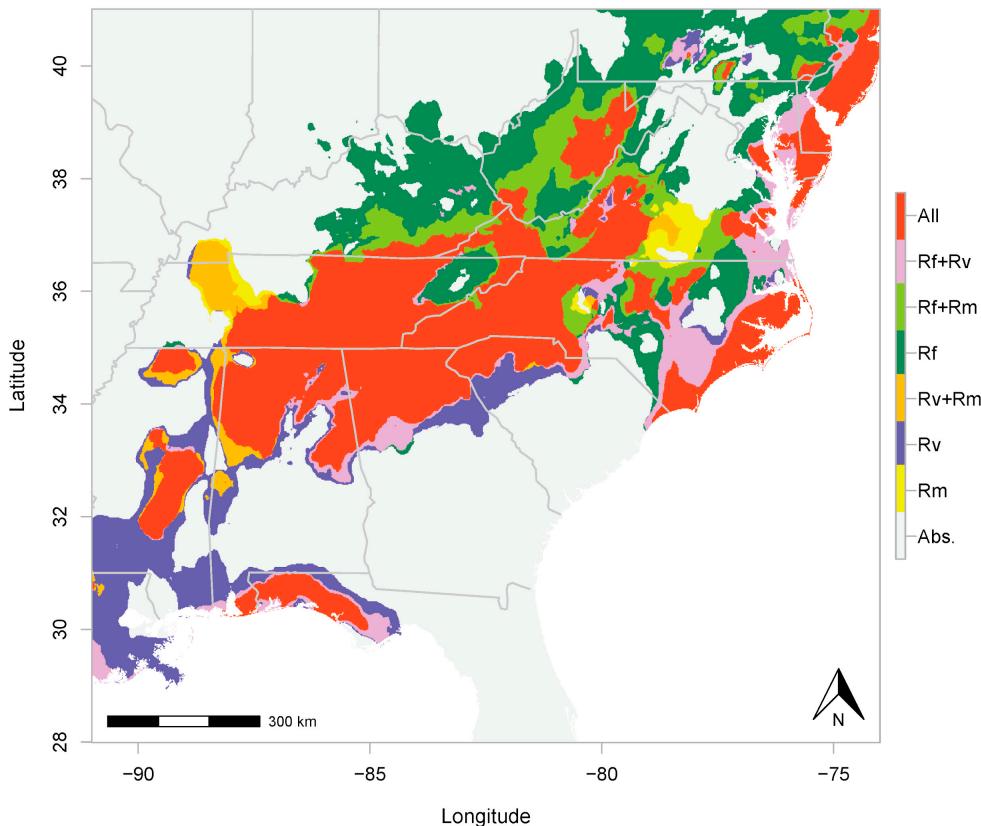
169 The predicted distribution of *R. flavipes* spanned a larger area in the northern portion of the
170 southern Appalachians than that of the other two species, overlapping with *R. virginicus* along the
171 Atlantic Coast, and with *R. mallei* in Kentucky, West Virginia, Virginia, and further north in
172 Pennsylvania (Figure 2). Predicted distributions of all three species overlapped in eastern Tennessee,
173 western North Carolina, northern Alabama and Georgia (Figure 2).



174
 175 **Figure 1.** Predicted niche occupancy. Four environmental factors were used to estimate niche occupancy
 176 of *R. flavipes* (Rf), *R. malletei* (Rm), and *R. virginicus* (Rv): left two panels: temperature range and
 177 summer temperature; right two panels: dry- and wet-season precipitation.

178 **Table 1.** Niche identity test. The upper off-diagonal shows Schoener's D statistic, and the lower off-
 179 diagonals shows the modified Hellinger statistic, I. Significant niche divergence is reported in bold
 180 text with red highlighting. The more dissimilar of the other two niche comparisons is highlighted in
 181 pink. Abbreviations used for *R. flavipes*, *R. malletei*, and *R. virginicus* are Rf, Rm, and Rv, respectively.

	Rf	Rm	Rv
Rf	-	D=0.744 P=0.280	D=0.582 P<0.001
	I=0.935 P=0.239	-	D=0.788 P=0.630
Rv	I=0.843 P<0.001	I=0.961 P=0.750	-

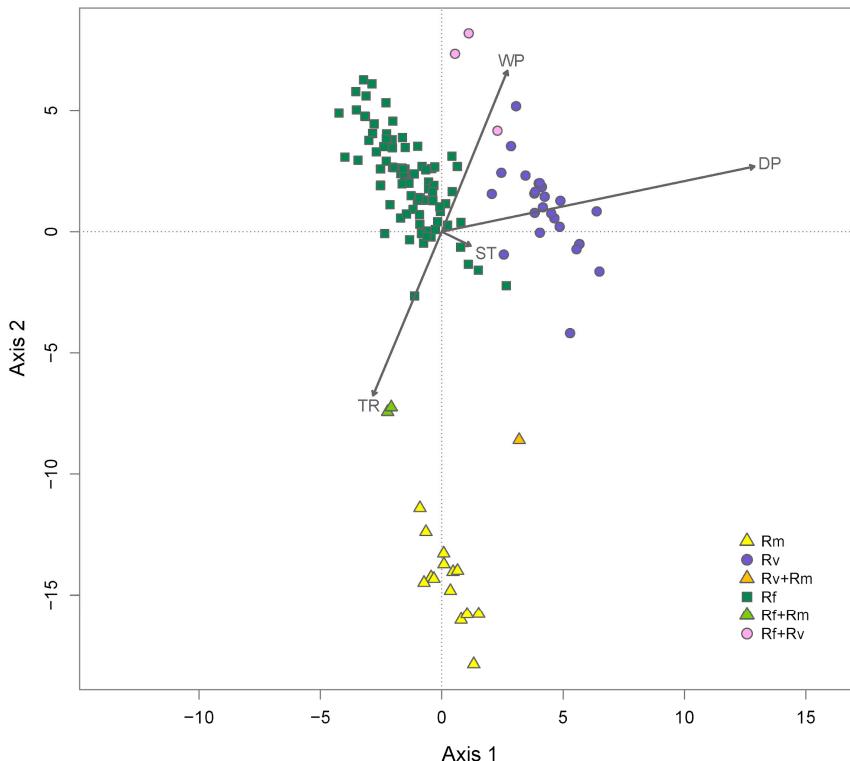


182
 183 **Figure 2.** Distributional overlap of *R. flavipes* (*Rf*), *R. malletei* (*Rm*), and *R. virginicus* (*Rv*). Overlap is
 184 color coded based on the number of species. “All” is where occurrence of all three species is predicted.
 185 Areas of two-species overlap are shown in the legend as “*Rf+Rv*,” “*Rf+Rm*,” and “*Rv+Rm*.” Absence
 186 of all three species is shown in grey and referred to in the legend as “Abs.”.

187 **3.2. Environmental factors and niche divergence**

188 Distance-based redundancy analysis (Figure 3) showed that only the dry-season precipitation
 189 factor contributed significantly ($F_{1, 127} = 8.673, P = 0.001$) to differences in occurrence among the three
 190 *Reticulitermes* species. After accounting for spatial structure by partialling out six significant spatial
 191 components (PCNM axes 1, 4, 6, 17, 43, and 58), dry-season precipitation remained significant ($F_{1, 121} = 5.622, P = 0.003$). The six significant spatial components along with dry-season precipitation
 192 accounted for 18.5% of the observed variation in the occurrence data. Spatial structure alone
 193 explained 9.6% of the variation, environmental factors accounted for 3.3%, and the interaction
 194 between the two explained an additional 5.6% of the variation.
 195

196 Following the removal of spatial structure effects, the highest correlation coefficient between
 197 environmental factors and ordination axes of the distance-based redundancy analysis was recorded
 198 for dry-season precipitation ($r = 0.730$) and axis 1. This axis captured the divergence of *R. virginicus*
 199 from the other two species (Figure 3). Thus, dry-season precipitation contributed significantly to *R. virginicus*
 200 divergence. While not significant, temperature range ($r = -0.383$) and wet-quarter
 201 precipitation ($r = 0.376$) were correlated with axis 2, which captured the divergence of *R. malletei*
 202 (Figure 3).



203

204 **Figure 3.** Distance-based redundancy analysis. The plot shows a constrained ordination of 132 sampling
 205 sites, color coded based on the number of species present. Sites where only *R. flavipes*, *R. virginicus*, or
 206 *R. mallei* were sampled are referred to in the legend as "Rf," "Rv," and "Rm," respectively. Two-
 207 species sites are shown in the legend as "Rf+Rv," "Rf+Rm," and "Rv+Rm." The ordination is
 208 conditional on six significant spatial components (PCNM axes 1, 4, 6, 17, 43, and 58) and constrained
 209 by four environmental factors: dry-season precipitation (DP); wet-season precipitation (WP); summer
 210 temperature (ST); temperature range (TR). Arrows show strength of correlation of environmental
 211 factors with ordination axes 1 and 2.

212 4. Discussion

213 This study provides insights into the ecology of subterranean termites with regard to geographic
 214 distributions and niche partitioning among three broadly co-distributed *Reticulitermes* species in the
 215 southern Appalachians Mountains and surrounding areas. This region is a biogeographically
 216 significant center of endemism, yet the ecology of its resident invertebrate fauna—particularly
 217 saproxylic insects—is poorly known. Our ENMs suggest that an area in the mid-latitudes of the
 218 southern Appalachians, characterized by complex topography and an array of six ecoregions,
 219 provides suitable habitat to support all three *Reticulitermes* species. Our study also highlights the roles
 220 that temperature and precipitation play in driving niche divergence among *Reticulitermes* species. To
 221 our knowledge, this work represents the first evidence of significant regional-scale niche divergence
 222 between *R. flavipes* and *R. virginicus*. Below, we consider the broader context of these findings, as well
 223 as caveats and future directions for follow-up studies that build on the information presented here.

224 4.1. *Reticulitermes* distributions and climatic drivers of niche divergence among species

225 Our analyses predicted extensive co-occurrence of all three *Reticulitermes* species in the mid-
 226 latitudes of the southern Appalachians (Figure 2). Based on paleoclimatic [45], biogeographic [46]
 227 and comparative phylogeographic [47] data, the southern Appalachians remained free from
 228 Pleistocene ice sheets and served as a major refuge for many species during glacial periods,
 229 consequently maintaining higher levels of biodiversity. Indeed, the present-day complexity of this
 230 mid-latitude region harbors many different niches, which could facilitate stable coexistence of closely
 231 related species. However, in addition to pronounced co-occurrence of *Reticulitermes* species in the

232 montane regions of the southern Appalachians, our ENMs also identified areas of two- and three-
233 species co-occurrence along the Gulf coast of western Florida, and the Atlantic coast from North
234 Carolina to New Jersey and New York. The Florida Gulf coast has previously been identified as an
235 important refuge for a forest-dependent millipede during the Last Glacial Maximum [48]. Indeed, the
236 paleoclimatic history of areas to the south and east of the southern Appalachian Mountains are
237 increasingly being recognized as reservoirs of forest invertebrate biodiversity during past periods of
238 environment change. The incidence of high termite species diversity—even though only assessed
239 here for one genus—is therefore not unexpected.

240 Our study provides additional insights into co-occurrence of *Reticulitermes* species. Consistent
241 with the findings of Maynard et al. [35], we determined that *R. flavipes* has a broad latitudinal range,
242 but *R. virginicus* is more restricted to the south. Maynard et al.'s [35] ENMs showed that temperature
243 variables were the most important predictors of termite distributions. Based on our formal
244 assessment of niche overlap between *R. flavipes* and *R. virginicus*, we determined that both
245 temperature and precipitation seasonality (as represented by temperature range, summer
246 temperature, and dry- and wet-season precipitation) play non-negligible roles in the significant niche
247 divergence between *R. flavipes* and *R. virginicus*. Furthermore, using distance-based redundancy
248 analysis, we identified dry-season precipitation as a major driver of this divergence.

249 4.2. Potential explanations for lack of empirical evidence for local-scale coexistence of *Reticulitermes* species

250 Interestingly, despite the significant niche divergence between *R. flavipes* and *R. virginicus*, we
251 collected both of these species from the same rotting log at one sampling site (i.e., #37 located near
252 the Georgia/Southern Carolina state border; Table S1). To our knowledge, this is the first record of
253 true syntopy between *Reticulitermes* species. The apparent rarity of syntopy and general lack of
254 coexistence of *Reticulitermes* species at local scales could be explained by competitive exclusion.
255 Colony size and soldier number are important features for termite competitive ability. Termite
256 species with small colonies have been observed to relinquish resources and be eliminated by
257 dominant interspecific competitors with large colonies [49]. Through avoidance of dominant
258 competitors, interspecific competition may result in spatial separation [50], but also temporal
259 separation (i.e., phenological differences). Termites may be able to avoid other related species using
260 vibrational cues. Indeed, vibrational cues are important for termite sensory perception and
261 communication, as these signals can travel over long distances [51,52]. For instance, the drywood
262 termite *Cryptotermes secundus* can distinguish conspecifics from the dominant competitor in the
263 environment, the subterranean termite *Coptotermes acinaciformis* [51]. Furthermore, *Coptotermes*
264 *acinaciformis* detects its major predator, the ant *Iridomyrmex purpureus*, using vibrational cues only
265 [52]. Overall, given these highly tuned sensory capabilities, it stands to reason that competitive
266 exclusion, or competitor avoidance, could be important factors in preventing local co-occurrence
267 among *Reticulitermes* species.

268 4.3. Caveats and future directions

269 While our sampling suggests that true syntopy and local co-occurrence of different species at
270 the same site is very rare, our detection of only one species in all but one rotting log, and at the
271 majority of sampling sites (i.e., 126 out of 132), may actually be a consequence of the sampling strategy
272 that was employed (see Methods). Briefly, we simply aimed to collect termites from each site, rather
273 than provide a complete assessment of termite diversity at each site. Indeed, variance partitioning
274 reflects this, showing that most (81.5%) of the variance in the occurrence data did not stem from
275 spatial structure (9.6%), or environmental differences (3.3%), or interaction between the two (5.6%).
276 Accordingly, while competitive exclusion is a plausible explanation for apparent rare local-scale co-
277 occurrence (i.e., micro-allopatry) among *Reticulitermes* species, a dedicated sampling approach would
278 be required to formally test this idea. For example, exhaustively sampling multiple logs per site, at a
279 series of sites arranged along a transect traversing a region where two or more species occur in close
280 proximity would be a productive approach. Fortunately, the present study identified specific

281 geographic areas where future assessments of the frequency of true syntopy vs. micro-allopatry, and
282 associated interspecific competitive interactions, should be focused (Table S1; Figure S1).

283 Although we have shown separation in niche space between species, particularly *R. flavipes* and
284 *R. virginicus*, these inferences were underpinned by regional-scale environmental variables, and so
285 they do not take into account local-scale drivers of niche divergence such as differences in
286 microhabitat preference, phenology, or diet. Indeed, Maynard et al. [35] highlighted that biotic and
287 soil characteristics play a role in termite distribution and abundance. Thus, our assessment of niche
288 divergence is necessarily incomplete. While it does provide important baseline information, follow-
289 up studies of local-scale drivers of species' distributions could examine aspects of the microhabitat
290 (e.g., humidity and temperature of soil and rotting logs), timing of nuptial flights along latitudinal
291 and altitudinal clines, and/or use stable isotopes to determine decomposition stage of ingested wood
292 and the importance of microbial biomass in termite diets at a given location [53].

293 **Supplementary Materials:** The following are available online, Supplementary File 1: Environmental variables
294 and Ecological Niche Modeling methods; Supplementary File 2: Figure S1: Map of *Reticulitermes* sampling
295 depicting occurrences of one or more species at each site; Supplementary File 3: Table S1: Sampling sites with
296 number of species occurrences at each site and number of logs per site; Supplementary File 4: Table S2. Pairwise
297 niche overlap among *Reticulitermes* species for each of four environmental factors.

298 **Author Contributions:** Both authors conceived the study, and collected and curated samples; C.H. performed
299 lab work, analyzed the data, and wrote the first draft; both authors edited subsequent drafts.

300 **Funding:** This work was supported by grants from the American Philosophical Society, Bay and Paul
301 Foundations, Eppley Foundation for Research, National Geographic Society, Washington Biologists' Field Club,
302 and start-up funds from the University of Mississippi.

303 **Acknowledgments:** We thank R.J. Dyer, R.E. Symula, and E.W. Collier for their assistance with field work, and
304 B.C. Collins for his contribution to lab work. Scientific collecting permits were provided by the Alabama
305 Department of Conservation and Natural Resources, Georgia Department of Natural Resources (permit number
306 29-WBH-12-16), United States Department of Agriculture Forest Service, and United States National Park Service
307 (permit numbers GRSM-2012-SCI-2242, SHEN-2012-SCI-0015, and CUGA-2012-SCI-0008).

308 **Conflicts of Interest:** The authors declare no conflict of interest.

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