
Quantifying Forest Cover Loss during the COVID-19 Pandemic in the Charcoal Production Basin of Lubumbashi (DR Congo) through Remote Sensing and Landscape Analysis

[Yannick Useni Sikuzani](#)*, [Médard Mpanda Mukenza](#)*, [Ildephonse Kipili Mwenya](#), [Héritier Khoji Muteya](#), [Dieu-donné N'tambwe Nghonda](#), [Nathan Kasanda Mukendi](#), François Malaisse, Françoise Malonga Kaj, Donatien Dibwe Dia Mwembu, [Jan Bogaert](#)

Posted Date: 30 April 2024

doi: 10.20944/preprints202404.1944.v1

Keywords: Forest cover loss, COVID-19 pandemic, Remote sensing, Landscape analysis, Miombo woodland, Sustainable forestry



Preprints.org is a free multidiscipline platform providing preprint service that is dedicated to making early versions of research outputs permanently available and citable. Preprints posted at Preprints.org appear in Web of Science, Crossref, Google Scholar, Scilit, Europe PMC.

Copyright: This is an open access article distributed under the Creative Commons Attribution License which permits unrestricted use, distribution, and reproduction in any medium, provided the original work is properly cited.

Article

Quantifying Forest Cover Loss during the COVID-19 Pandemic in the Charcoal Production Basin of Lubumbashi (DR Congo) through Remote Sensing and Landscape Analysis

Yannick Useni Sikuzani ^{1,*}, Médard Mpanda Mukenza ^{2,*}, Ildephonse Kipili Mwenya ², Hérítier Khoji Muteya ¹, Dieu-donné N'tambwe Nghonda ¹, Nathan Kasanda Mukendi ³, François Malaisse ⁴, Françoise Malonga Kaj ⁵, Donatien Dibwe Dia Mwembu ⁶ and Jan Bogaert ⁴

¹ Unité Ecologie, Restauration Ecologique et Paysage, Faculté des Sciences Agronomiques, Université de Lubumbashi, Lubumbashi P.O. Box 1825, Democratic Republic of the Congo; hkhoji@doct.uliege.be (H.K.M.); nghondan@unilu.ac.cd (D.-d.N.N.)

² Département de Gestion des Ressources Naturelles Renouvelables, Faculté des Sciences Agronomiques, Université Technologique Katumba Mwanke, Lubumbashi P.O. Box 74, Democratic Republic of the Congo; ildephonsekipilim@gmail.com

³ Unité de recherche en Economie et développement agricole, Faculté des Sciences Agronomiques, Université de Lubumbashi, 1825 Lubumbashi, RD Congo ; kasanda.mukendi@unilu.ac.cd (N.K.M.)

⁴ Unité Biodiversité et Paysage, Gembloux Agro-Bio Tech, Université de Liège, 5030 Gembloux, Belgium; malaisse1234@gmail.com (F.M.); j.bogaert@uliege.be (J.B.)

⁵ Ecole de santé publique, Université de Lubumbashi, Lubumbashi P.O. Box 1825, Democratic Republic of the Congo; kaj_mal@Unilu.ac.cd (F.M.K.)

⁶ Département des Sciences Historiques, Faculté des Lettres et Sciences Humaines, Lubumbashi P.O. Box 1825, Democratic Republic of the Congo; dibwe.mwembu@unilu.ac.cd (D.D.D.M.)

* Correspondence: sikuzaniu@unilu.ac.cd (Y.U.S.); mpandamedard@gmail.com (M.M.M.)

Abstract: In the context of the Lubumbashi Coal Production Basin (LCPB), the socio-economic repercussions of the COVID-19 pandemic have exacerbated pressures on populations dependent on forest resources for their subsistence. This study employs a comprehensive methodological approach, integrating advanced remote sensing techniques, including image classification, mapping, and detailed landscape analysis, to quantify alterations in forest cover within the LCPB during the pandemic period. Our findings reveal a consistent trend of declining forested area, characterized by processes of attrition and fragmentation observed throughout various study phases, spanning from May 2019 to November 2023. This reduction in forest cover, notably more pronounced in the vicinity of Lubumbashi city and the northern zone of the LCPB, proved to be less pronounced between November 2019 and September 2020, underscoring the influence of COVID-19 pandemic-induced confinement measures on forest management practices in the region. However, subsequent to this period of restriction, deforestation activity intensified, leading to significant landscape transformations within the LCPB, primarily attributable to expanded human activities, consequently resulting in a notable decrease in the proportion of land occupied by these natural ecosystems. Consequently, the size of the largest forest patch declined substantially, decreasing from 14.62% to 8.20% between May 2019 and November 2023, thereby fostering a heightened density of forest edges over time. Our findings provide a significant contribution to understanding the complex interactions between the COVID-19 pandemic and deforestation phenomena, emphasizing the urgent need to adopt adaptive management strategies and appropriate conservation measures in response to current economic challenges.

Keywords: forest cover loss; COVID-19 pandemic; remote sensing; landscape analysis; miombo woodland; sustainable forestry

1. Introduction

Forest ecosystems play a pivotal and multifaceted role on a global scale. Their primary contribution to climate regulation stems from their ability to sequester carbon dioxide (CO₂) [1], serving as critical carbon sinks and significantly mitigating greenhouse gas emissions [2]. Concurrently, forests uphold the stability of hydrological cycles by regulating water flow dynamics. Moreover, they harbor unparalleled biodiversity, safeguarding the planet's genetic diversity [3]. Additionally, forests provide indispensable resources such as timber, firewood, and traditional medicines, thereby contributing to the daily welfare of populations [2,4,5]. Notably, an estimated 880 million people worldwide, particularly women, rely on firewood collection or charcoal production, underscoring forests' indispensable role in human livelihoods [6,7].

Nevertheless, the conservation and sustainable management of forests remain imperative for preserving ecological equilibrium, fostering biodiversity, and supporting local communities [8]. Between 1990 and 2020, the global forest area diminished by 178 million hectares, equivalent to the size of Libya. Over 100 million hectares of forests face various threats such as fires, pests, diseases, invasive species, droughts, and adverse climatic events [9,10]. Agricultural expansion remains the primary driver of deforestation, forest fragmentation, and biodiversity loss, particularly in tropical regions [11].

Furthermore, scholars suggest that the COVID-19 pandemic could have enduring economic and social repercussions with ongoing implications for forests [12,13]. For instance, crises such as COVID-19 might precipitate an escalation in forest fires due to reduced monitoring capacities [13]. Moreover, the surge in urban unemployment resulting from the pandemic has prompted temporary migration to rural areas, where individuals seek livelihood opportunities such as farming in proximity to forests, thereby exacerbating pressure on these ecosystems [14,15]. Conversely, studies indicate that the initial impacts of COVID-19 might have led to a reduction in deforestation rates due to economic slowdown during containment measures [16].

While Latin America halved the rate of forest loss from 5.2 million hectares per year between 2000 and 2010 to 2.6 million hectares per year over the subsequent decade, forest loss in Africa increased from 3.4 to 3.9 million hectares per year [17]. In Africa, deforestation is closely linked to small-scale activities such as subsistence farming, charcoal production, and firewood collection [18]. The Democratic Republic of the Congo (DR Congo) significantly contributes to this deforestation, becoming the country with the second-fastest rate of forest loss after Brazil [9].

In the southeastern DR Congo, the Katangese Copperbelt Area emerges as a longstanding mining hub, playing a central role in demographic growth [19]. Concurrently, this region is traversed by the *miombo* forest, an open forest formation characterized by sparse shrub vegetation, dominated by trees of the *Brachystegia*, *Julbernardia*, and *Isoberlinia* genera [20,21]. The regression of the *miombo* forest is particularly concentrated around mining areas, with the Lubumbashi region serving as a notable example. Driven by rapid urbanization, deforestation is exacerbated by unsustainable logging to meet local energy, construction, and other needs, as well as agricultural practices like shifting cultivation and farmland expansion [22–24]. As a significant mining locale, Lubumbashi also contributes to deforestation through the expansion of mining operations and associated infrastructure. Challenges in forest resource management, such as the issuance of forest concessions without adequate oversight by customary authorities [25,26], have further exacerbated deforestation, particularly during the COVID-19 pandemic, in the rural area of Lubumbashi known as the Lubumbashi Charcoal Production Basin (LCPB).

At the outset of the COVID-19 pandemic, Africa, including the DR Congo and the LCPB, implemented measures such as lockdowns to contain the virus's spread [27]. These restrictions aimed to safeguard the population and bolster healthcare system capacities [28]. Within this basin, the containment measures and economic disruptions associated with the COVID-19 pandemic have intensified pressure on populations reliant on forest resources for their livelihoods [23,26,29]. Some have responded by escalating logging activities to meet immediate economic needs, as exemplified in Kasomeno within the same region with the illicit exploitation of *Pterocarpus tinctorius* Wel. [25]. Similarly, population groups affected by economic disruptions or containment measures may have

migrated to forested areas, thereby exerting additional pressure on these fragile ecosystems. Pandemic-related restrictions have also altered electricity consumption patterns, particularly with increased demand in residential areas of Lubumbashi due to remote work. This heightened demand for electricity has disrupted energy supply, exacerbating the demand for charcoal [30]. Food security has become a major concern, prompting inhabitants to clear new areas for agriculture. Additionally, during pandemics, roads have facilitated access to forested areas, intensifying pressure on these ecosystems. Reduced surveillance and controls have fostered unsustainable utilization of forest resources, including logging and land conversion for agricultural purposes, along extensive distances to villages and roads. Changes in population movement patterns due to travel restrictions and economic impacts have led to displacement to forested areas, exacerbating deforestation, particularly near roads and villages. The pandemic has also influenced the demand for forest products, such as timber and non-timber products, with roads facilitating the transport of these products to markets, and shifts in demand potentially affecting deforestation rates, as observed in Peru [14].

Analyzing spatial landscape patterns and their dynamics provides valuable insights into fundamental ecological processes, and vice versa [31]. In this context, remote sensing plays a crucial role in monitoring deforestation by offering accurate, continuous, and large-scale data [32,33]. Additionally, landscape ecology tools furnish an integrated spatial approach to comprehending deforestation dynamics. By amalgamating advanced geospatial methods with landscape ecology tools, detailed information on changes in forest landscapes' processes is furnished, thereby enabling more informed and sustainable forest resource management [34]. This study aims to assess and quantitatively analyze the impacts of COVID-19 on deforestation using a comprehensive approach integrating remote sensing and landscape analysis. We hypothesize that the socio-economic constraints imposed by containment measures during the COVID-19 pandemic would likely have led to a substantial surge in human activities during subsequent periods, thereby contributing to a significant decline in forest cover. This decline is anticipated to be more pronounced in peripheral areas of the city and in the northern region of the landscape, particularly attributable to the dense network of paved roads facilitating increased human access and activity.

2. Materials and Methods

2.1. Study Area

This investigation was conducted within the LCPB, situated in the southeastern sector of the DR Congo. Geographically, the area is delineated by coordinates spanning from 10°39'7.47"-12°26'37.61" S and 26°20'54.95"-28°40'13.55" E (Figure 1) [30]. Encompassing a total land area of approximately 26,603.4 km², the LCPB features a climate categorized as Cw according to the Köppen classification system [35]. Seasonally, the region experiences a rainy period extending from November to March, followed by a dry phase from May to September, interspersed with transitional periods in April and October [20,36]. The average annual precipitation in this area amounts to around 1,270 mm, with a mean annual temperature of 20°C, although recent climate fluctuations have been noted [36]. The prevailing vegetation, *miombo* woodland, is presently fragmented and intermittently replaced by savannah terrain [30,37]. The dominant soil is ferralitic according to FAO/ World Reference Base for Soil Resources classification, with an average pH level of approximately 5.2 [38]. The socioeconomic activities within the LCPB encompass a range of sectors including mining, agriculture, charcoal production, livestock husbandry, and commerce, all of which hold substantial significance for the local population exceeding 3.5 million inhabitants [39,40]. The region faces notable challenges, including rapid population growth and deficiencies in public services and infrastructures. However, since the appointment of provincial governors in 2006, the construction of roads linking Lubumbashi to various territories has led to the proliferation and establishment of settlements within the LCPB. Nevertheless, the accelerated population expansion, coupled with persistent poverty, has resulted in the emergence of numerous villages within the LCPB [30]. Consequently, residents heavily depend on natural resources to meet essential needs and fulfill the demands of urban dwellers for agricultural produce, non-timber forest products, and charcoal [23,26]. This economic reliance has exerted

significant pressure on forest resources, contributing to notable deforestation and forest degradation in the region. Additionally, deficiencies and inconsistencies in forest governance exacerbate the situation by allowing for unsustainable exploitation of forest resources [26].

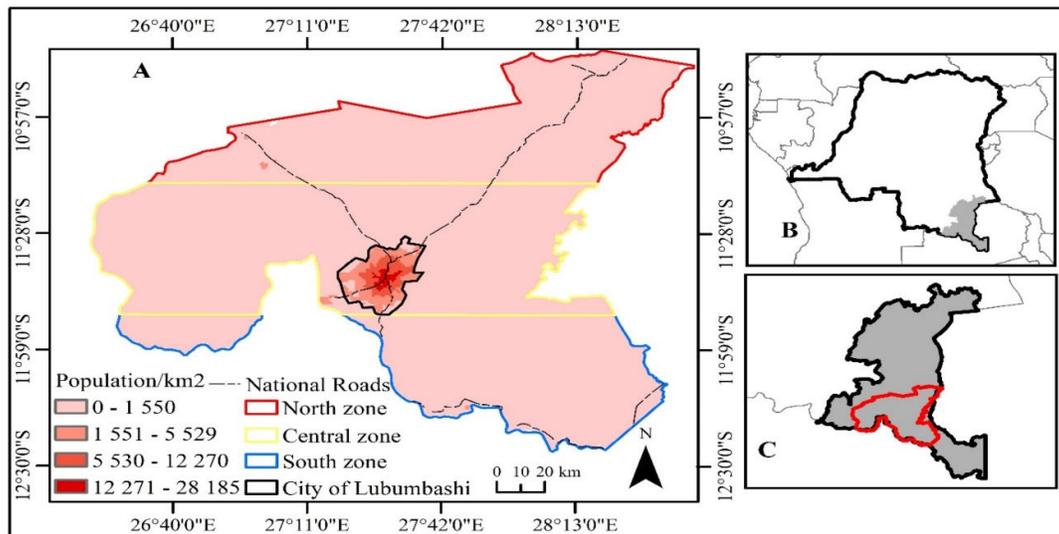


Figure 1. Location of the study area (A), the Charcoal Production Basin (LCPB), in southeastern DR Congo (B) within the Upper Katanga province (C). For this study, the LCPB was subdivided into three parts: the northern zone (~8000 km²), the central zone (~12000 km²) and the southern zone (~7000 km²). Population density estimation was obtained from the WorldPop website, based at the University of Southampton. The data presented were generated in 2020.

2.2. Data

Data from the Sentinel-2 satellite, operating at a 10-meter resolution and capturing imagery during the dry season, was pivotal in analyzing landscape changes. This satellite system, developed by the European Space Agency (ESA) as part of the Copernicus program, features a multispectral sensor that records data across different spectral bands [41]. Its high-resolution imagery provided valuable information crucial for the study's analysis of landscape dynamics and transformations [42–44]. The investigation employed a series of Sentinel-2 images acquired at intervals spanning from 2019 to 2023, encompassing periods preceding, during, and after the emergence of the COVID-19 pandemic, with the objective of conducting a comprehensive evaluation of temporal variations. These time points delineate three discrete phases: the pre-COVID-19 era (May and November 2019), the concurrent pandemic phase (July 2020, September 2020, May 2021, May 2022), and the post-pandemic era (November 2022, May 2023, and November 2023). It is pertinent to mention that significant public health interventions were implemented during the second phase coinciding with the COVID-19 pandemic. Specifically, a state of health emergency, along with containment measures, was enforced from 24 March 2020 to 21 July 2020, aligning with the time frame of satellite image acquisition from November 2019 to July 2020. Subsequently, a curfew without a state of health emergency was instituted from 21 December 2020 to 15 February 2021, corresponding to image captures between November 2020 and May 2021. The selection of diverse dates within this investigation also facilitates an assessment of the impact of a project on community management of *miombo* forests in southeastern Katanga, a collaborative effort between the FAO and the Government of the DR Congo. This initiative, active from 1 August 2016 to 31 March 2023, was sanctioned by the creation of community forest concessions within the LCPB from February 10, 2021.

2.3. Sentinel-2 Images Classification

The Sentinel-2 satellite imagery employed in this research underwent pre-processing procedures, including atmospheric correction and radiometric calibration, to enhance image quality and reliability, thus enabling subsequent analyses [41]. These corrections were automated using the Sensor-Invariant Atmospheric Correction (SIAC) module on the Google Earth Engine (GEE) platform, utilizing a sensor-invariant methodology to minimize sensor-specific dependencies [45]. A false-color composite of Sentinel-2 images was generated by combining near infrared, red and green bands to enhance vegetation discrimination [46]. This false-color composite was crucial in revealing detailed information within the images, facilitating nuanced interpretation of captured objects. For precise analysis, distinct land cover units were identified and coded across scenes, and training areas known as Regions of Interest (ROIs) were delineated in each study year, strategically selected during the dry season for temporal consistency [40]. These ROIs, consisting of polygons with 2 and 4 pixels per area, were tailored for each land cover type, totaling 200 ROIs per type, aimed at reducing transitional effects known as the 'mixel' effect [47,48]. The ROIs were utilized as training data to construct a model based on the random forest algorithm, which combines predictions from multiple decision trees to enhance prediction accuracy and minimize classification errors [49,50]. The use of GEE was pivotal, providing cloud-based infrastructure for efficient processing of vast satellite datasets and utilizing substantial computational resources [51].

Land cover types, encompassing forests, savannas, built-up areas, bare soil, croplands, and water bodies, were grouped into two primary landscape classes: forests (*miombo* woodland, dry dense forest, gallery forest) and non-forests (all other land cover types). To evaluate classification accuracy, we followed best practices recommended by Olofsson et al. [52], employing unbiased surface estimators and estimating uncertainty through reference observations from change maps between 2019 and 2023. Samples were stratified based on stable and change classes for each period, and sample sizes were determined using Cochran's method [53]. A total of 11,595 points were sampled across eight periods, with proportions allocated based on stratum sizes. Error matrices, expressing estimated surface proportions and confidence intervals, were generated in QGIS 3.26.1 software, along with measurement accuracies such as overall accuracy, user accuracy, and producer accuracy. Land cover maps were produced using ArcGIS 10.8.

2.4. Land Cover Change Analysis

The research utilized Fragstats 4.2 software to assess the influence of human activities on landscape morphology within the LCPB by computing five landscape metrics. These metrics, encompassing class area, mean patch area, number of patches, largest patch index, and Edge density, clarified patch fragmentation across different land cover types [54,55]. Class area signifies the percentage of a specific land cover type within a given landscape. This measurement is valuable for understanding the composition of the landscape by pinpointing the predominant landscape matrix. Meanwhile, average patch area refers to the average size of habitat patches within a land cover type, offering insights into habitat fragmentation. Edge density quantifies the total length of all edge segments per hectare for a land cover type (m/ha). It gauges the complexity or ruggedness of the patches within a land cover types; high edge density indicates that ruggedness decreases as patches merge and boundaries dissolve, while it increases with the emergence of new fragments or patches. Another crucial metric is the number of patches within a land cover type, where a high patch number may indicate fragmentation and dispersion, while a low patch number suggests infilling and aggregation. Such metrics hold ecological importance as they provide valuable insights into spatial patterns and fragmentation within forest landscapes [56], aiding in evaluating habitat fragmentation levels, patch connectivity, and overall landscape health [55,57]. A monthly deforestation rate was computed to standardize the comparison of forest cover loss progression across different periods under examination. This deforestation rate is defined as the ratio between the difference in forest cover between two periods and the difference in months separating them.

Spatial transformation processes affecting land cover dynamics were identified using a decision tree algorithm [58]. The decision tree uses comparisons between the number of patches, total area,

and total perimeter of a land cover type to identify the corresponding spatial transformation process in the landscape between two specific dates [58]. Between two specific dates, a decrease in patch number, coupled with reduced class area, signifies attrition. Conversely, increased class area indicates aggregation. Unaffected patch numbers alongside increased class area suggest enlargement, while increased class area and patch number indicate creation of new patches. A decrease in class area alongside increased patch number, with minimal area loss due to linear disruptions, denotes dissection. Fragmentation combines patch increase with substantial class area loss. A decrease in class area can lead to perforation if the total perimeter increases, or patch shrinkage if it remains constant. In cases where patch number and class area remain constant, a constant total perimeter suggests shift, while a changing total perimeter indicates deformation [58]. To distinguish between fragmentation and dissection processes, the ratio of total areas at different time points was analyzed, with a ratio above 0.75 indicating dissection and a ratio at or below 0.75 indicating fragmentation prevalence [59].

The landscape composition analysis within the Lubumbashi Conservation Pilot Project (LCPB) was systematically conducted across its northern, central, and southern zones. The Kolmogorov-Smirnov test was applied to evaluate the conformity of the data to a normal distribution. Given the data's adherence to normality, an analysis of variance (ANOVA) was utilized to compare forest proportions across these zones, thereby assessing the impact of zone classification. Moreover, to investigate the variation in forested areas surrounding Lubumbashi, ArcGIS 10.8 software was utilized to create five buffer zones ranging from 0 to 150 kilometers (0-30km, 30-60km, 60-90km, 90-120km, and 120-150km) from the city center (refer to Figure 2). This facilitated an examination of the impact of proximity to the city on forest cover loss. The Kruskal-Wallis test, employed as a non-parametric alternative to ANOVA due to the non-normal distribution of the data, was employed to compare forest proportions at varying distances from the city of Lubumbashi.

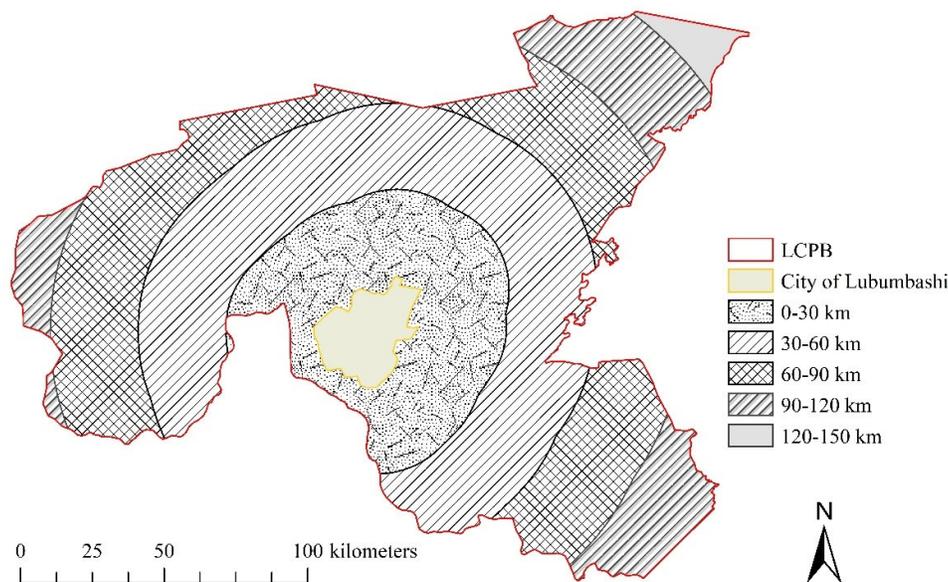


Figure 2. Evaluation device used to measure the effect of distance from the city of Lubumbashi on forest degradation and deforestation.

3. Results

3.1. Classification Accuracy

Table 1 below presents a synthesis of the accuracy results from supervised classifications of Sentinel-2 images obtained using the Random Forest classifier, covering the periods from May 2019 to November 2023. These results, derived from Sentinel-2 image reclassification, are detailed in the

appendix. It is important to note that the images were classified with an overall accuracy exceeding 90% for all periods studied, demonstrating a high reliability in distinguishing between different land cover types. The user's and producer's accuracy values ranging between 86% and 100% attest to the undeniable quality of the results obtained. Furthermore, the 95% confidence interval calculated to estimate the stratified area of each land cover class for all study periods remains below 5%, further reinforcing the reliability of the study's conclusions.

Table 1. Accuracy assessment and for the land cover and land cover change map based on the Sentinel-2 image supervised classification using the Random Forest classifier.

May 2019_November 2019	Forest	Forest gain	Forest loss	Non-forest
<i>Accuracy measure</i>				
Prod. acc.	100.0%	87.0%	100.0%	100.0%
User acc.	98.7%	100.0%	100.0%	100.0%
Overall acc.	99.3%			
November 2019-July 2020	Forest	Forest gain	Forest loss	Non-forest
Prod. acc.	100.0%	100.0%	97.0%	100.0%
User acc.	99.6%	100.0%	100.0%	100.0%
Overall acc.	99.8%			
July 2020-September 2020	Forest	Forest gain	Forest loss	Non-forest
Prod. acc.	100.0%	98.8%	94.1%	100.0%
User acc.	98.8%	100.0%	100.0%	99.8%
Overall acc.	99.4%			
September 2020_May 2021	Forest	Forest gain	Forest loss	Non-forest
Prod. acc.	100.0%	100.0%	99.2%	100.0%
User acc.	99.8%	100.0%	100.0%	100.0%
Overall acc.	99.9%			
May 2021-May 2022	Forest	Forest gain	Forest loss	Non-forest
Prod. acc.	100.0%	100.0%	100.0%	100.0%
User acc.	100.0%	100.0%	100.0%	100.0%
Overall acc.	100.0%			
May 2022-November 2022	Forest	Forest gain	Forest loss	Non-forest
Prod. acc.	100.0%	100.0%	100.0%	100.0%
User acc.	100.0%	100.0%	100.0%	100.0%
Overall acc.	100.0%			
November 2022-May 2023	Forest	Forest gain	Forest loss	Non-forest
Prod. acc.	100.0%	98.1%	92.5%	100.0%
User acc.	99.8%	100.0%	100.0%	99.0%
Overall acc.	99.4%			
May 2023-November 2023	Forest	Forest gain	Forest loss	Non-forest
Prod. acc.	100.0%	100.0%	100.0%	100.0%
User acc.	100.0%	100.0%	100.0%	100.0%
Overall acc.	100.0%			

3.2. Land Cover Change Mapping

The visual analysis of land cover maps presented in Figure 3 highlights changes in the landscape between May 2019 and November 2023. Specifically, a pronounced forest cover loss is observed, primarily in the northern part of the study area, while non-forest areas continue to expand in all directions. These changes are even more noticeable between November 2022 and November 2023, a period following the onset of the COVID-19 pandemic. These results suggest that the COVID-19 pandemic may have had an impact on landscape spatial pattern changes through the deforestation process within the LCPB.

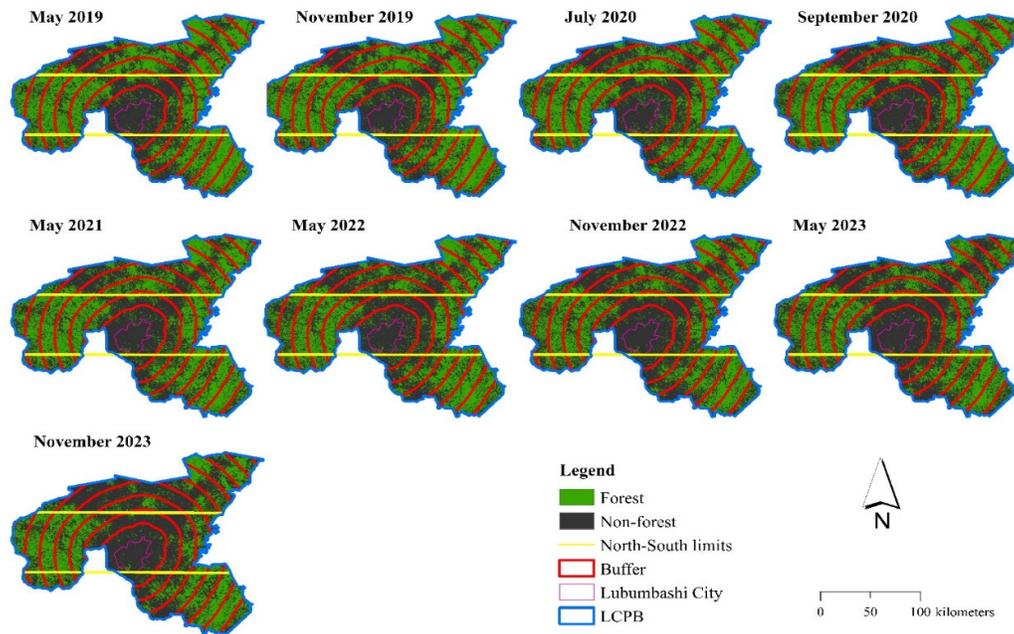


Figure 3. Mapping of the evolution of land cover within the LCPB for May 2019, November 2019, July 2020, September 2020, May 2021, May 2022, November 2022, May 2023 and November 2023, obtained through the “Random Forest” classifier.

3.3. Landscape Composition Dynamics

During the period from May 2019 to November 2023, a gradual forest cover loss was observed across the entire LCPB (Figure 4). This decrease occurred in favor of non-forest. These results demonstrate a significant transition wherein the forest, which represented approximately 52% of the landscape in May 2019, was replaced by non-forest areas that became the dominant matrix after July 2020. By November 2023, forest cover had fallen below 40% of the landscape, experiencing a loss of approximately one-quarter compared to its initial area in May 2019 (Figure 4a). However, our results reveal significant variations in monthly deforestation rates over time (Figure 4b). Between May 2019 and November 2019, the monthly deforestation rate was relatively low at 0.05%. This period was followed by a slight decrease in the rate between November 2019 and July 2020, to 0.02%. However, a notable increase in the deforestation rate was observed between July 2020 and September 2020, reaching 0.59%. This period was marked by a significant rise in deforestation activities. Subsequently, although the deforestation rate decreased between September 2020 and May 2021 (0.43%), it remained relatively high compared to the earlier study periods. This downward trend continued until May 2022, with a deforestation rate of 0.24%. However, from May 2022 to November 2022, a new increase in the deforestation rate was recorded, reaching 0.46%. Between November 2022 and May 2023, although the deforestation rate slightly decreased to 0.44%, it remained higher than the levels observed in the earlier study periods. Finally, the period from May 2023 to November 2023 saw another increase in the deforestation rate, reaching 0.52%. These results highlight the complexity of deforestation dynamics within the LCPB (Figure 4b).

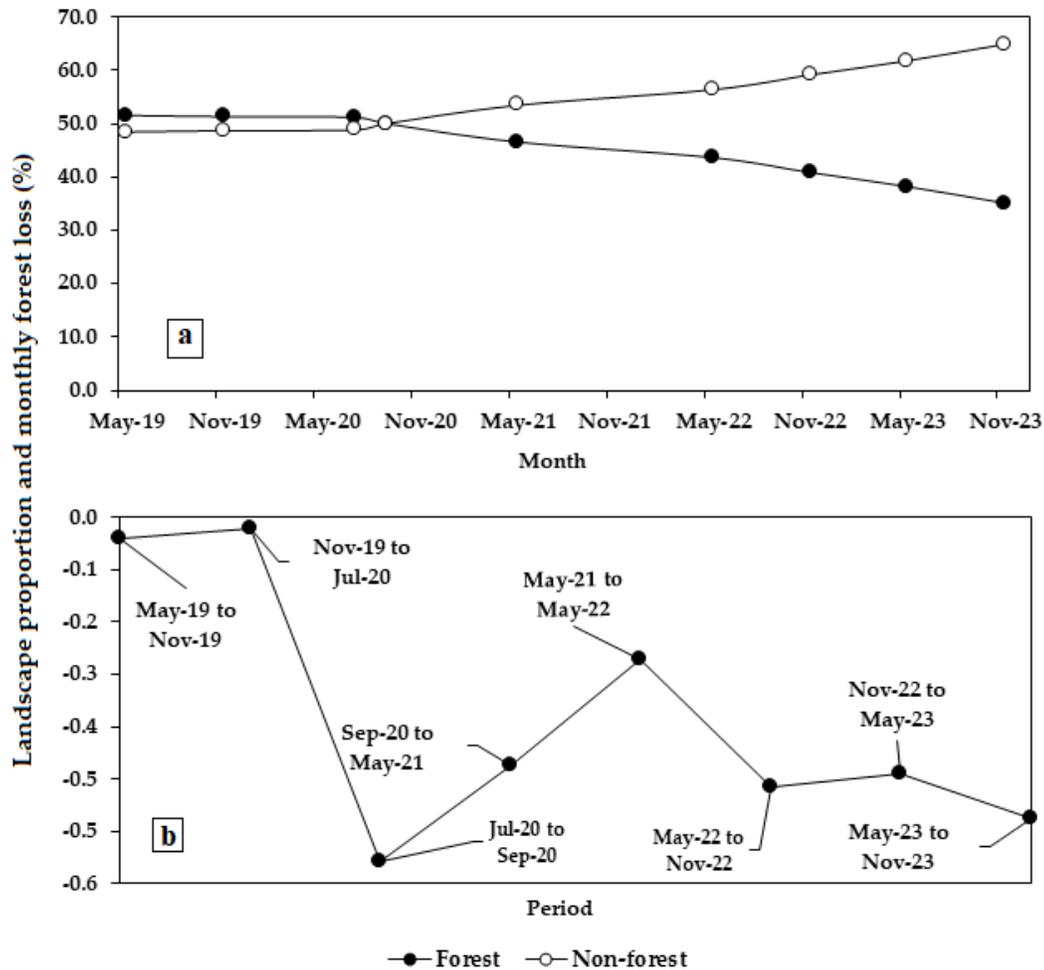


Figure 4. Temporal dynamics of forest cover changes in the LCPB: evolution of forest cover in the LCPB between May 2019 and November 2023 (a); evolution of the monthly forest loss (b).

Analyses conducted over different key periods of the study regarding the evolution of forest cover proportion revealed statistically significant differences ($p = 0.0027$). Before the COVID-19 pandemic, forests constituted the main component of the landscape in the LCPB (~52%). However, during the COVID-19 pandemic, the proportion of forests significantly decreased, ceasing to be the predominant landscape component (~48%). After the pandemic, a substantial gap widened further between forest and non-forest occupations. Forests experienced a significant decrease during this period, reaching a proportion of less than 40% within the LCPB landscape. These results underscore the importance of monitoring and protecting forests in the context of biodiversity and current environmental challenges. They also highlight the potential impact of global events, such as the COVID-19 pandemic, on local natural ecosystems.

The spatial analysis of forest cover around the city of Lubumbashi reveals significant trends ($p=0.015$) and notable variations (Figure 5). The results show a high proportion of forests over longer distances from the city, proportionally to the distance from the city center. Overall, a consistent decrease in forest cover is observed over time, notably on the outskirts of Lubumbashi city. However, an exception is noted between May 2019 and July 2020, where forest cover remained relatively stable regardless of the distance from the city. During the COVID-19 pandemic lockdown period, between November 2019 and July 2020, a significant drop in forest area was recorded far from the city of Lubumbashi, particularly in areas beyond 120 km from the city. Three distinct trends are evident. Firstly, forest cover, while showing a steady decline, continues to dominate the land within the 60-

150km radius of the city. Secondly, in the 30-60 km radius, forest cover declined to nearly 40% by November 2023, contrasting with an approximately 58% proportion in May 2019. Lastly, within the area less than 30km from the city, forest cover, accounting for slightly under 20% of the land area in May 2019, had diminished to nearly 10% by November 2023. This analysis suggests that proximity to Lubumbashi has an impact on surrounding forest density, and specific events such as lockdowns can lead to major variations in forest cover on the periphery.

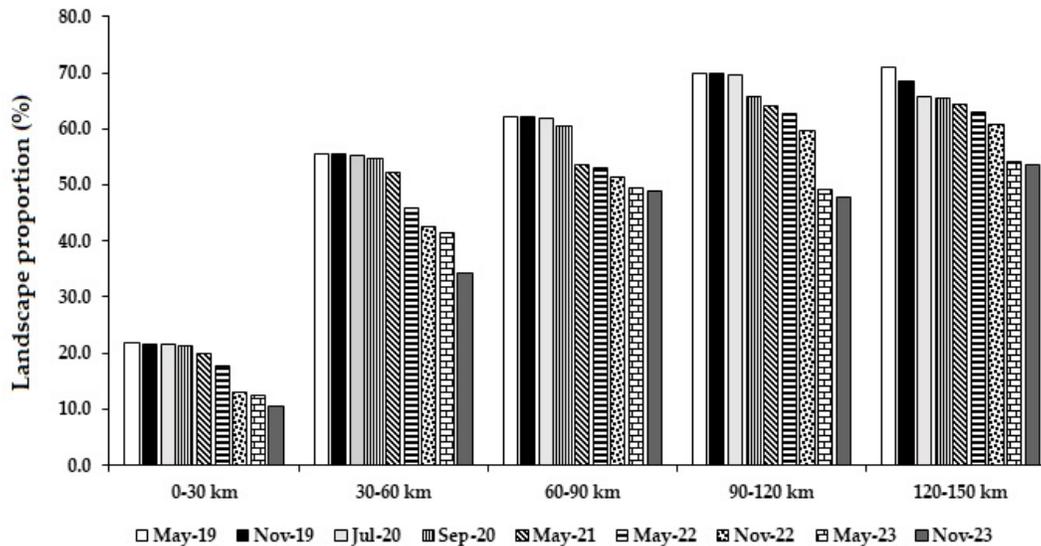


Figure 5. The variation in forest area (%) based on distance from the city of Lubumbashi between November 2019 and November 2023. Distance from Lubumbashi plays a crucial role in the deforestation in the LCPB.

The gradual forest cover loss observed at the LCPB scale was also noted in its three distinct zones: the Northern Zone (NZ), the Central Zone (CZ), and the Southern Zone (SZ) (Figure 6). Globally, the results reveal a consistent decrease in the proportion of forested areas across the three zones of Lubumbashi city, over the studied period from May 2019 to November 2023. In the Northern zone of the city of Lubumbashi, forest cover declined from 54% in May 2019 to 28% in November 2023, indicating a steady and substantial reduction over time. This significant decrease signifies a noteworthy loss of forested areas in this region. Similarly, in the Southern zone of the city of Lubumbashi, the proportion of forest cover decreased from 55% in May 2019 to 44% in November 2023. While the reduction in forest cover is less pronounced compared to the Northern zone, this consistent downward trend also underscores environmental degradation in this area. In the Central zone surrounding the city of Lubumbashi, forest cover also experienced a decline, dropping from 48% in May 2019 to 34% in November 2023. Although this decrease is slightly less pronounced than in the Northern and Southern zones, it nevertheless exhibits a linear downward trend over the study period. These findings highlight a significant degradation of forest across all zones during the study period. This reduction in forest cover can have significant implications for biodiversity, ecosystem services, and the resilience of local ecosystems in the face of environmental changes. However, prior to the COVID-19 pandemic (from May to November 2019) and during the period covering the initial lockdown (from November 2019 to July 2020), a low deforestation dynamic was observed in the NZ, CZ, and SZ compared to other periods. Analysis of land cover change between May 2019 and November 2023 revealed growing concerns regarding forest loss, particularly in the Northern and Central Zones where the forest proportion significantly decreased ($p < 0.05$). These results suggest that deforestation has intensified over time, especially in the Northern and Central Zones, raising significant concerns about biodiversity loss and environmental degradation in this region.

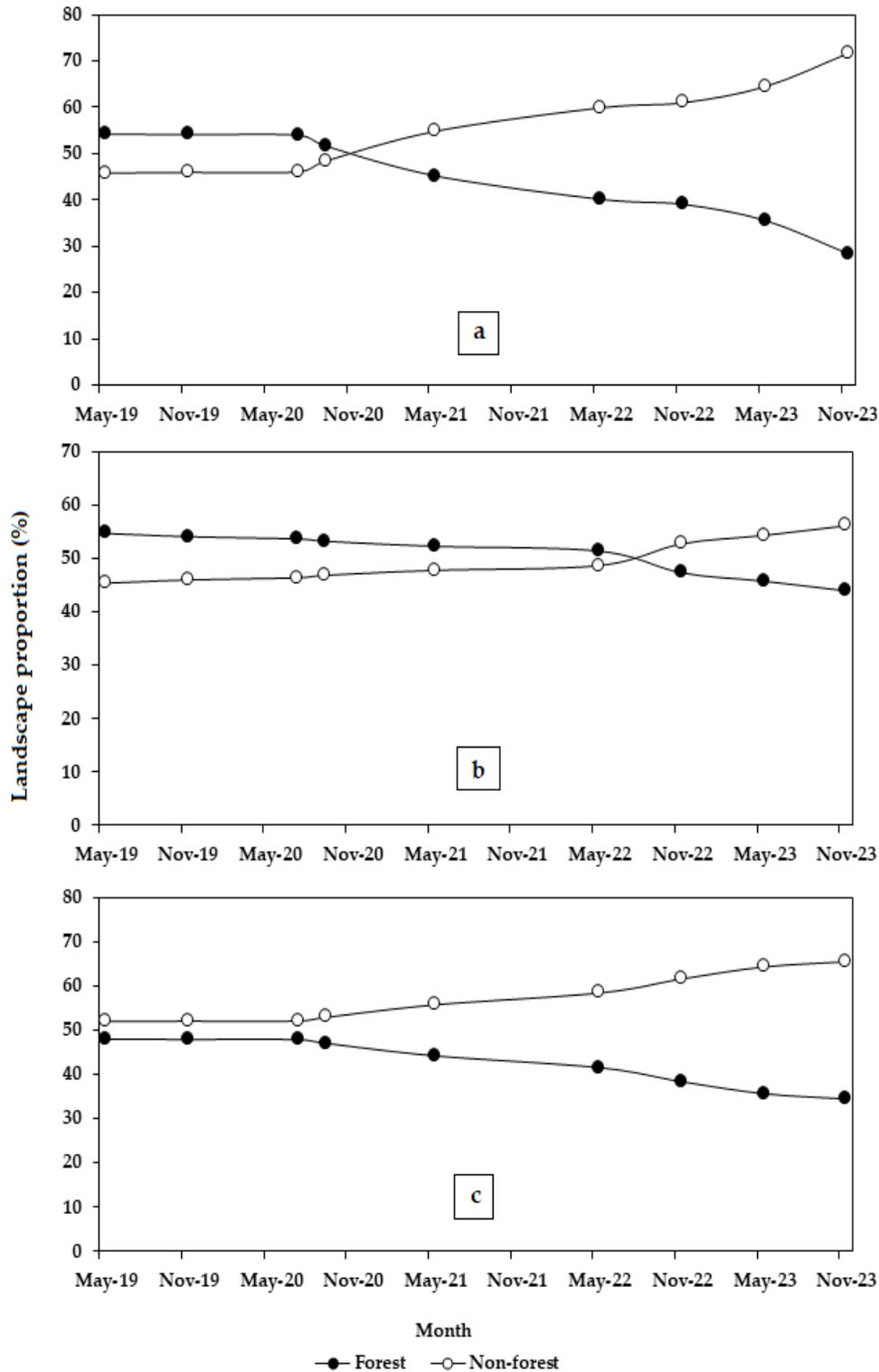


Figure 6. Evolution of forest cover between November 2019 and November 2023 in the 3 zones of the Lubumbashi charcoal production basin (a=North zone, b=South zone and c= central zone).

3.4. Change in the Configuration of Landscape

The analysis of the landscape reveals distinct spatial transformation processes over the study periods, as presented in Table 2. From May 2019 to November 2019, November 2019 to July 2020, and

from May 2023 to November 2023, there was a notable decrease in both the number of forest patches and their respective class area, signifying a spatial transformation process of attrition in spatial configuration. Conversely, the periods spanning from July 2020 to May 2023 are marked by the dissection of forest patches (with a t-value exceeding 0.75) as dominant spatial transformation process since the decrease in class area is accompanied by an increase in patch number. Over the period from May 2019 to November 2023, small forest patches gradually dominated larger ones, leading to a continuous reduction in the average area of forest zones. Consequently, over the years, the forest has significantly degraded. This degradation is highlighted by a decrease in the Largest Patch Index (LPI) values (13% in May 2019 to 8% in November 2023) in parallel to the increase in the edge density values between May 2019 and May 2023 (87 to 108), before dropping to 86 in November 2023. These results underscore the complex changes and pressures exerted on the forest during the study period, with potentially significant implications for biodiversity and local ecosystems.

Table 1. Land cover class configuration indices in November 2019, July 2020, September 2020, May 2021, May 2022, November 2022, May 2023 and November 2023. LPI: largest patch index (%), CA: class area (km²), MA: mean patch area (km²), NP: number of patches, ED: Edge density (m/ha).

Date	NP	CA	MA	LPI	ED
May-2019	400858.0	13711.4	6.8	14.6	87.3
November-2019	733771.0	13654.4	6.3	13.6	88.7
July-2020	662269.0	13613.0	6.9	15.6	84.8
September-2020	1526810.0	13297.6	4.5	13.2	116.6
May-2021	869037.0	12388.0	6.2	11.4	89.2
May-2022	1057747.0	11607.7	4.9	10.1	89.5
November-2022	1171325.0	10865.8	4.1	10.4	98.7
May-2023	1220557.0	10160.1	3.4	9.8	108.3
November-2023	1168247.0	9849.8	3.7	8.2	85.7

4. Discussion

4.1. Methodology

Traditional aerial imagery, although valuable, presents limitations such as cost and insufficient coverage over extended periods when compared to satellite imagery. This constraint impedes the comprehensive analysis of temporal changes [61]. Conversely, the utilization of Sentinel-type satellite images in this study provides high spatial resolution, facilitating detailed examination of land cover alterations that may not be adequately captured by medium to coarse resolution imagery, thus enabling precise analysis [62,63]. To ensure dependable classification outcomes, multiple sets of image datasets were acquired at various intervals, enabling the comparison and assessment of changes pre- and post-COVID-19 pandemic. This temporal methodology offers historical and current insights into forest cover dynamics [64]. Despite challenges in synchronizing Sentinel-2 image acquisition dates with COVID-19 pandemic-related events, mitigating seasonal influences by capturing images during specific dry season periods facilitated the comparison of deforestation data between pre-pandemic and pandemic periods. Integration of remote sensing data with ground observations validated the findings, enhancing the robustness of the analysis. This integrated approach allowed for the isolation of the specific impacts of COVID-19 on deforestation [65].

Our investigation employs rigorous scientific methodologies to examine the ramifications of COVID-19 on deforestation. Specifically, preprocessing steps are crucial for eliminating distortions and enhancing data quality, thereby improving result reliability [41]. Geometric rectification ensures accurate comparisons across different timeframes by eliminating geometric variations and ensuring precise interpretation of forest cover changes [41]. The use of a supervised classification approach enables accurate discrimination between land cover types, aiding in the identification of landscape composition changes with high precision compared to unsupervised classification methods [23]. Additionally, change detection analysis to identify forest cover loss involved calculating a series of

landscape metrics [55,66]. Indeed, the conversion of natural landscapes into anthropized landscapes results in significant spatial pattern alterations. Employing landscape metrics allows for quantitative analysis of changes, providing detailed information on spatial patterns and forest fragmentation. The complexity of landscape pattern studies is evident in the multitude of metrics utilized, as a singular value cannot fully encapsulate landscape configuration [34]. Understanding forest spatial transformation processes involves analyzing patch number, class area, and perimeter, forming the basis of decision-making processes such as the decision tree, which integrates information on patch morphology variation and landscape transformation causes.

4.2. Forest Cover Loss during Covid-19 Pandemic: Drivers, Extent, and Spatio-Temporal Dynamics

In the Lubumbashi region, deforestation rates were high, but they decreased during the lockdown period before rising again post-restriction. This trend continued until the official end of the pandemic in July 2023. The lockdown and restrictions of the global COVID-19 pandemic resulted in significant environmental changes and reduced anthropogenic disturbances on ecosystems worldwide, known as the "anthropause" [67]. Activities like unauthorized logging and artisanal mining were reduced during the lockdown, leading to a decrease in deforestation levels in the LCPB.

During the lockdown, non-compliance with COVID-19 measures likely led to undetected infections, impacting people's ability to exploit forest resources. This could affect food and economic security in areas reliant on forest resources. The pandemic exacerbated fragility in forest resource monitoring due to reduced field presence of public service agents and limited resources [68,69]. Economic impacts and limited compliance led to increased forest cover loss, as many turned to informal activities over official duties. The population below the poverty line resorted to exploiting forest resources during the curfew, leading to increased deforestation rates. After lockdown, reorienting phytomedicines for COVID-19 treatment reduced the risk of weakened populations relying on forest resources [70].

The DR Congo encountered the onset of viral infectious diseases like Ebola nearly a decade ago, yet national initiatives to fortify the health system's resilience against health hazards and calamities remained absent. Subsequently, following the relaxation of lockdown measures and the upsurge in human activities, deforestation accelerated rapidly. This surge is attributed to heightened demands on forest resources due to the recommencement of economic operations and the relaxation of travel constraints, which facilitated the resurgence of human activities, including illicit ones [71]. However, during the COVID-19 pandemic, there was a notable surge in the utilization of traditional medicine. The heightened demand for medicinal plants utilized in traditional medicine placed additional stress on forest reserves [72], with communities intensifying their collection efforts [73]. Furthermore, the commercialization of medicinal plants flourished during the pandemic, incentivizing collectors to search for more specimens in forested areas [74–78].

A trend in the decrease in the extent of forest cover loss was observed after Covid-19 curfew period, probably due to the implementation of the community forestry project. This project has identified within the LCPB 30 sites for local forest concessions covering 334,656 hectares and submitted 20 applications approved by the governorate. It has also promoted agroforestry on 20,882 hectares, encouraging sustainable management of forest resources and reducing pressure on deforested areas [79]. Community-based Forest management has been identified as a win-win option for reducing deforestation while improving the well-being of rural communities [80,81]. Unfortunately, the following period was characterized by an increase in deforestation extent attributable to increased human activity in forest concessions and agroforestry practices, possibly due to economic opportunities and population growth.

Our research reveals a trend of escalating forest exploitation in areas increasingly distant from Lubumbashi city. As forests diminish in nearby rural zones proximate to the city, demand for residual forest resources such as timber, firewood, non-timber forest products, and agricultural land escalates significantly [26]. This heightened demand leads to unsustainable forest exploitation and the conversion of remote forested lands into agricultural zones. Forest scarcity consequently amplifies the fragmentation of remaining forested areas, driving farmers to seek new lands for conversion,

including the remaining forest areas, as observed in Ghana and Burkina Faso [82]. Moreover, deforestation itself can perpetuate forest scarcity by diminishing the size and quality of remaining forest areas, compromising their natural regeneration capacity and complicating forest ecosystem restoration efforts, fostering a cycle of continual deforestation [83]. While previous studies reported a deforestation radius of 35 km [84], Kabulu et al.'s analysis based on 2008 field data revealed that charcoal production for Lubumbashi extended up to approximately 60 km from the city [85]. Our findings suggest this radius has expanded to over 100 km, propelled by urban growth and population expansion, escalating land and forest resource demands. Advancements in technology facilitate access to remote forested areas, while inadequate regulations foster uncontrolled exploitation, particularly in remote rural zones governed by customary chiefs. Notably, state authority diminishes in rural areas away from urban centers, favoring customary chiefs' influence [86]. The alarming deforestation near Lubumbashi during the COVID-19 pandemic stems from various factors. Urban expansion in Lubumbashi has exerted substantial pressure on local forests, driven by the escalating demand for land for residential, industrial, and agricultural purposes [87]. Furthermore, global economic pressures reinforced the phenomenon, with increased demands for agricultural and forest products compelling companies to expand at the expense of forests. In the Kasenga territory, in response to illegal exploitation of *Pterocarpus tinctorius*, forest cover experienced a significant decline between 2009 and 2021, evidenced by the replacement of forest matrix with savanna in 2021 [25].

The region north of the studied landscape, intersected by two major asphalt roads (national roads to the West and 5 to the East), witnessed substantial deforestation during the COVID-19 pandemic. The presence of asphalt roads is acknowledged to enhance access to forested regions, leading to heightened resource utilization, particularly evident in the post-lockdown phase [88]. The pandemic's disruptions to traditional economic pursuits potentially spurred some individuals to seek alternative livelihoods such as logging or agriculture, especially in locales accessible via asphalt roads. Moreover, the infrastructure provided by asphalt roads facilitated economic activities and the establishment of transient farmer or charcoal producer settlements, amplifying land demand [89]. Notably, the Lubumbashi-Likasi and Lubumbashi-Kasenga Road networks, situated in the northern region of our study area, are primary conduits for transporting charcoal to Lubumbashi [90]. In contrast, the central and southern areas, lacking significant asphalt roads, are more distant from populous centers and markets, thereby diminishing economic incentives for exploitation. However, recent research indicates that ongoing road network expansions, encompassing the southern region supplying charcoal to Lubumbashi, are anticipated to reduce existing forest resources by over 2% while increasing agricultural lands by about 16% [91].

Acknowledging the various stages of human-induced landscape transformation, encompassing perforation, dissection, and fragmentation [92], our observations within the LCPB are in line with this trend. In our study, forest degradation predominantly stems from dissection and attrition, aligning with previous conclusions [25,93]. Forest fragmentation often precedes the disappearance of certain fragments and the reduction in the size of forest patches over time due to various factors [94]. Firstly, the expansion of human activities leads to the division of forests into smaller pieces [94]. Subsequently, these isolated fragments face increased pressure, resulting in their gradual disappearance [95]. Additionally, interventions such as selective logging can cause a reduction in the size of remaining patches, explaining the decrease in the LPI over time [96]. Finally, Human activities such as selective deforestation and unsustainable logging lead to an increase in the proportion of forest edges [97]. This occurs because these activities create clearings around forest edges, promoting fragmentation and loss of forest continuity. Consequences include increased vulnerability of ecosystems to biological invasions, habitat alteration for wildlife, and reduced resilience to disturbances.

4.2. Implications for Forest Management

The reduction in forest cover leads to habitat loss and fragmentation, resulting in a decline in biodiversity within the *miombo* woodland [98]. Strict conservation measures, such as protected area designation and sustainable land management practices, can help preserve critical habitats and

mitigate biodiversity loss. Moreover, decreasing forest cover diminishes the *miombo* woodland's capacity to provide essential ecosystem services, including carbon sequestration, water regulation, and soil stabilization [99]. Promoting reforestation and afforestation initiatives can restore ecosystem functions and enhance the provision of vital ecosystem services crucial for local communities and ecosystem resilience [100].

Forest degradation also impacts the livelihoods and well-being of communities reliant on forest resources for food, fuel, medicine, and traditional practices [101]. Implementing community-based forest management approaches that empower local communities to sustainably manage forest resources can foster socio-economic development while conserving biodiversity [102]. Deforestation contributes significantly to climate change by releasing stored carbon into the atmosphere and reducing forests' capacity to sequester carbon dioxide [103]. Enhancing climate change mitigation strategies through reforestation, afforestation, and sustainable land use practices can help mitigate climate change impacts and enhance ecosystem resilience [104]. Forest cover loss can also alter hydrological patterns, affecting water availability, quality, and regulating functions [105]. Implementing watershed management strategies, such as reforestation of riparian zones and conservation of water catchment areas, can improve water resource management and ensure long-term water security [106]. Additionally, deforestation increases soil erosion, leading to soil degradation, loss of fertility, and increased sedimentation in water bodies. Implementing agroforestry practices, such as tree planting on degraded lands, can prevent soil erosion, improve soil health, and enhance agricultural productivity sustainably [107].

Inadequate enforcement of forest conservation policies, weak governance, and lack of stakeholder engagement contribute to unsustainable practices of forest exploitation [108]. Strengthening policy frameworks, enhancing law enforcement, promoting stakeholder participation, and fostering partnerships between government, NGOs, and local communities can improve forest governance and promote sustainable land management [109]. Moreover, limited awareness and understanding of the importance of forest conservation and sustainable land management practices among local communities and stakeholders hinder conservation efforts [110]. Investing in environmental education, raising awareness about the value of forests, and promoting community participation in conservation activities can foster a culture of environmental stewardship and support sustainable land management practices.

Furthermore, the reduction in forest cover can lead to habitat loss for wildlife, potentially increasing their contact with humans and domestic animals, thus facilitating the spillover of zoonotic diseases from wildlife to humans [111]. Conservation efforts should focus on protecting and restoring forested areas, implementing sustainable land-use practices [112], and creating wildlife corridors to reduce habitat fragmentation [113]. Indeed, the spatial process of attrition and dissection of forest patches can lead to isolated forest fragments, which may not support viable wildlife populations or ecological functions [113]. Land-use planning should prioritize connectivity between forest patches, promoting landscape-scale conservation strategies and incorporating ecological corridors into development plans to mitigate this [114]. The reduction in the size of the largest forest patch increases edge effects, leading to habitat degradation and potential changes in species composition. Restoring and protecting large contiguous forest areas, along with implementing measures to reduce edge effects like buffer zones and wildlife-friendly land-use practices, should be prioritized [115].

Finally, pandemics or epidemics exacerbate poverty, leading to a cascade of events that intensify the exploitation of forest resources, thus fueling a cycle of disease emergence [116]. Firstly, health crises cause widespread job loss and income reduction, particularly in countries lacking robust social safety nets. Impacted communities often turn to forests for sustenance, driving unsustainable practices like deforestation, wood extraction, hunting, and gathering non-timber products [117]. Secondly, disrupted food supply chains heighten food insecurity, prompting reliance on forests for immediate needs like food, medicine, and construction materials, further straining resources [118]. Thirdly, weakened governance during health crises fosters illegal exploitation of forests, including wildlife trade, increasing the risk of zoonotic disease transmission. These pressures alter ecosystems, disturb wildlife habitats, and heighten human-wildlife interactions, facilitating pathogen

transmission and fostering new disease outbreaks [119]. To disrupt this cycle, multifaceted interventions are necessary, addressing socioeconomic vulnerabilities, promoting sustainable practices, and enhancing governance to safeguard both human health and forest ecosystems [120,121].

5. Conclusions

By combining landscape ecology analysis tools with remote sensing and geographic information systems (GIS), this study analyzed landscape changes in the Lubumbashi Charcoal Production Basin (LCPB) from May 2019 to November 2023. Due to rapid population growth and economic disruptions caused by the COVID-19 pandemic, the study area experienced significant spatial changes, notably a continuous decrease in forest cover. However, this reduction was slower between November 2019 and September 2020, during the initial phase of COVID-19 pandemic-related lockdowns. This period also witnessed a substantial loss of forests on the outskirts of Lubumbashi city. Attrition and dissection of forest patches led to a landscape imbalance, paving the way for alternative land cover that expanded their footprint. Consequently, the level of landscape spatial pattern and composition changes intensified, indicating the growing impact of human activity on natural ecosystems. Finally, our results revealed that forest cover loss is notably more pronounced in the vicinity of Lubumbashi city and the northern zone of the LCPB. These results underscore the urgent need for conservation measures and sustainable natural resource management in the region to preserve biodiversity and ensure the long-term sustainability of forest ecosystems. They highlight the necessity of adopting effective land management and planning policies.

6. Patents

Author Contributions: Y.U.S.: conceptualisation, methodology and writing—original draft preparation; M.M.M.: conceptualisation, methodology and writing—original draft preparation, I.M.M.: data curation; H.K.M., D.-d.N.N., N.K.M., F.M.K., D.D.D.M. & F.M.: writing—review and editing; J.B.: supervision, writing—original draft and funding acquisition. All authors have read and agreed to the published version of the manuscript.

Funding: This was funded by the program ARES-CCD (Belgium).

Data Availability Statement: All the data presented in this study are available in this manuscript.

Acknowledgments: The authors would like to thank the development research project "Capacity building for the sustainable management of the miombo clear forest through the assessment of the environmental impact of charcoal production and the improvement of forest resource management practices (CHARLU)" and all the customary authorities for the facilities provided.

Conflicts of Interest: The authors state that the research was conducted without any commercial or financial relationships that could be interpreted as a potential conflict of interest.

References

1. Brockerhoff, E.G.; Barbaro, L.; Castagneyrol, B.; Forrester, D.I.; Gardiner, B.; González-Olabarria, J.R.; Lyver, P.O.B.; Meurisse, N.; Oxbrough, A.; Taki, H.; Thompson, I.D. Forest biodiversity, ecosystem functioning and the provision of ecosystem services. *Biodivers. Conserv.* **2017**, *26*, 3005-3035. <https://doi.org/10.1007/s10531-017-1453-2>.
2. Cooper, L.; MacFarlane, D. Climate-Smart Forestry: Promise and risks for forests, society, and climate. *PLOS Climate* **2023**, *2*, e0000212. <https://doi.org/10.1371/journal.pclm.0000212>.
3. Sukara, E. Tropical Forest Biodiversity to Provide Food, Health and Energy Solution of the Rapid Growth of Modern Society. *Procedia Environ Sci* **2014**, *20*, 803–8. <https://doi.org/10.1016/j.proenv.2014.03.097>.
4. Useni, S.Y.; Malaisse, F.; Yona, J.M.; Mwamba, T.M.; Bogaert, J. Diversity, use and management of household-located fruit trees in two rapidly developing towns in Southeastern D.R. Congo. *Urban For. Urban Green.* **2021**, *63*, 127220. <https://doi.org/10.1016/j.ufug.2021.127220>.
5. Friedlingstein, P.; O'sullivan, M.; Jones, M.W.; Andrew, R.M.; Gregor, L.; Hauck, J.; Le Quéré, C.; Luijkx, I.T.; Olsen, A.; Peters, G.P.; Peters, W. Global carbon budget 2022. *Earth Syst. Sci. Data Discuss.*, **2022**, pp.1-159. <https://doi.org/10.5194/essd-14-4811-2022>.

6. UN: United Nations Department of Economic and Social Affairs, 2021. The Global Forest Goals Report 2021. Sales No.: E. 21. IV. 3. Available online : <https://www.un.org/en/desa/global-forest-goals-report-2021> (accessed on 10 March 2024).
7. FAO & UNEP. The State of the World's Forests 2020. Forests, biodiversity and people. Rome. DOI: 10.4060/ca8642en. 2020. <https://doi.org/10.2307/208372>.
8. Vancutsem, C.; Achard, F.; Pekel, J.F.; Vieilledent, G.; Carboni, S.; Simonetti, D.; Gallego, J.; Aragao, L.E.; Nasi, R. Long-term (1990–2019) monitoring of forest cover changes in the humid tropics. *Sci. Adv.* **2001**, *7*, p.eabe1603. <https://doi.org/10.1126/sciadv.abe1603>.
9. FAO. Global Forest Resources Assessment (FRA), **2020**. <https://doi.org/10.1163/157180808X353939>.
10. Mulatu, K.A.; Mora, B.; Kooistra, L.; Herold, M. Biodiversity Monitoring in Changing Tropical Forests: A Review of Approaches and New Opportunities. *Remote Sens.* **2017**, *9*, 1059. <https://doi.org/10.3390/rs9101059>
11. Curtis, P.G., Slay, C.M., Harris, N.L., Tyukavina, A.; Hansen, M.C. Classifying drivers of global forest loss. *Science* **2018**, *361*(6407), 1108-1111. <https://doi.org/10.7163/PrzG.2022.4.4>.
12. Zahawi, R.A., Reid, J.L.; Fagan M.E. Potential impacts of COVID-19 on tropical forest recovery. *Biotropica* **2020**, *52*, 803–7. <https://doi.org/10.1111/btp.12851>.
13. López-Feldman, A.; Chávez, C.; Vélez, M.A.; Bejarano, H.; Chimeli, A.B.; Féres, J.; Robalino, J.; Salcedo, R.; Viteri, C. Environmental impacts and policy responses to Covid-19: a view from Latin America. *Environ. Resour. Econ.* **2020**, p.1. <https://doi.org/10.1007/s10640-020-00460-x>.
14. Céspedes, J.; Sylvester, J.M.; Pérez-Marulanda, L.; Paz-García, P.; Reymondin, L.; Khodadadi, M.; Tello, J.J.; Castro-Nunez, A. Has global deforestation accelerated due to the COVID-19 pandemic? *J. For. Res.* **2023**, *34*, pp.1153-1165. <https://doi.org/10.1007/s11676-022-01561-7>.
15. Couch, K.A.; Fairlie, R.W.; Xu, H.. Early evidence of the impacts of COVID-19 on minority unemployment. *Journal of public economics*, 2020, *192*, 104287. <https://doi.org/10.1016/j.jpubeco.2020.104287>
16. de Oliveira, P.C. Ethnophytomedicine in the Amazon: a Future Threatened by Deforestation and COVID-19 Pandemic. *Biomed J. Sci. Tech. Res.* **2021**, *33*. <https://doi.org/10.26717/bjstr.2021.33.005451>.
17. FAO. Latin American and Caribbean Forestry 2021. Available online: <https://www.fao.org/3/cb6027en/cb6027en.pdf> (accessed on 01 March 2024).
18. Eba'a Atyi, R.; Hiol Hiol, F.; Lescuyer, G.; Mayaux, P.; Defourny, P.; Bayol, N.; Nasi, R. Les forêts du Bassin du Congo: état des Forêts; CIFOR: Bogor, Indonésie, 2021. <https://doi.org/10.17528/cifor/008700>.
19. Potapov, P.V.; Turubanova, S.A.; Hansen, M.C.; Adusei, B.; Broich, M.; Altstt, A.; Mane, L.; Justice, C.O. Quantifying forest cover loss in Democratic Republic of the Congo, 2000–2010, with Landsat ETM+data. *Remote Sens. Environ.* **2012**, *122*, 106–116. <https://doi.org/10.1016/j.rse.2011.08.027>.
20. Malaisse, F. How to live and survive in Zambesian open forest (Miombo Ecoregion). Les presses agronomiques de Gembloux, Gembloux, Belgique, **2010**.
21. Useni, S.Y.; Malaisse, F.; Kaleba, S.C.; Kankumbi, F.M.; Bogaert, J. Le rayon de déforestation autour de la ville de Lubumbashi (Haut-Katanga, RD Congo): Synthèse. *Tropicicultura* **2017**, *35*, 215–221.
22. Useni, S.Y.; Muteya, H.K.; Bogaert, J. Miombo woodland, an ecosystem at risk of disappearance in the Lufira Biosphere Reserve (Upper Katanga, DR Congo)? A 39-years analysis based on Landsat images. *Glob. Ecol. Conserv.* **2020**, *24*, e01333. <https://doi.org/10.1016/j.gecco.2020.e01333>.
23. Cabala, K.S.; Sikuzani, Y.U.; Yamba, A.M.; Kankumbi, F.M.; Bogaert, J. Activités anthropiques et dynamique des écosystèmes forestiers dans les zones territoriales de l'Arc Cuprifère Katangais (RD Congo). *Tropicicultura* **2022**, *40*, 27. <https://doi.org/10.25518/2295-8010.2100>.
24. Khoji, M.K.; N'Tambwe, N.D.-D.; Malaisse, F.; Waselin, S.; Sambiéni, K.R.; Cabala, K.S.; Munyemba, K.F.; Bastin, J.-F.; Bogaert, J.; Useni, S.Y. Quantification and Simulation of Landscape Anthropization around the Mining Agglomerations of Southeastern Katanga (DR Congo) between 1979 and 2090. *Land* **2022**, *11*, 850. <https://doi.org/10.3390/land11060850>.
25. Mpanda, M.M.; Muteya, H.K.; Nghonda, D.-D.N.; Sambiéni, K.R.; Malaisse, F.; Kaleba, S.C.; Bogaert, J.; Sikuzani, Y.U. Uncontrolled Exploitation of *Pterocarpus tinctorius* Welw. and Associated Landscape Dynamics in the Kasenga Territory: Case of the Rural Area of Kasomeno (DR Congo). *Land* **2022**, *11*, 1541. <https://doi.org/10.3390/land11091541>.
26. N'tambwe Nghonda, D.-d.; Muteya, H.K.; Kashiki, B.K.W.N.; Sambiéni, K.R.; Malaisse, F.; Sikuzani, Y.U.; Kalenga, W.M.; Bogaert, J. Towards an Inclusive Approach to Forest Management: Highlight of the Perception and Participation of Local Communities in the Management of miombo Woodlands around Lubumbashi (Haut-Katanga, D.R. Congo). *Forests* **2023**, *14*, 687. <https://doi.org/10.3390/f14040687>.
27. Happi, C.T.; Nkengasong, J.N. Two years of COVID-19 in Africa: lessons for the world. *Nature* **2022**, *601*, 22–5.
28. Murewanhema, G.; Dzinamarira, T. The COVID-19 Pandemic: Public Health Responses in Sub-Saharan Africa. *Int. J. Environ. Res. Public Health* **2022**, *19*, 4448. <https://doi.org/10.3390/ijerph19084448>
29. Mukendi, N.K.; Muteya, H.K.; Nghonda, D.-d.N.; Berti, F.; Useni, S.Y.; Ndjibu, L.N.; Mbay, J.-P.K.; Fyama, J.N.M.; Lebailly, P.; Bogaert, J. Quantification and Determinants of Carbonization Yield in the Rural Zone

- of Lubumbashi, DR Congo: Implications for Sustainable Charcoal Production. *Forests* **2024**, *15*, 554. <https://doi.org/10.3390/f15030554>.
30. Khoji, M.H.; Nghonda, D.-d.N.; Kalenda, F.M.; Strammer, H.; Kankumbi, F.M.; Malaisse, F.; Bastin, J.-F.; Sikuzani, Y.U.; Bogaert, J. Mapping and Quantification of Miombo Deforestation in the Lubumbashi Charcoal Production Basin (DR Congo): Spatial Extent and Changes between 1990 and 2022. *Land* **2023**, *12*, 1852. <https://doi.org/10.3390/land12101852>.
 31. Richiardi, C.; Minciardi, M.R.; Siniscalco, C.; Adamo, M. Cumulative Spatial and Temporal Analysis of Anthropogenic Impacts in the Protected Area of the Gran Paradiso National Park in the NW Alps, Italy. *Land* **2023**, *12*, 1124. <https://doi.org/10.3390/land12061124>
 32. Shimabukuro, Y.E. ; Arai, E. ; Duarte, V. ; Jorge, A. ; Santos, E.G.D. ; Gasparini, K.A.C. ; Dutra, A.C. Monitoring deforestation and forest degradation using multi-temporal fraction images derived from Landsat sensor data in the Brazilian Amazon. *Int. J. Remote Sens.*, **2019**, *40*(14), 5475-5496. <https://doi.org/10.1080/01431161.2019.1579943>
 33. Foody, G.M. Remote sensing in landscape ecology. *Landscape Ecol.* **2023**, *38*;2711–6. <https://doi.org/10.1007/s10980-023-01753-4>.
 34. Zaehring, J.G.; Hett, C.; Ramamonjisoa, B.; Messerli, P. Beyond deforestation monitoring in conservation hotspots: Analysing landscape dynamics in north-eastern Madagascar. *Appl. Geogr.*, **2016**, *68*, 9-19. <https://doi.org/10.1016/j.apgeog.2015.12.009>
 35. Kotteck, M.; Grieser, J.; Beck, C.; Rudolf, B.; Rubel, F. World map of the Köppen-Geiger climate classification updated. *Meteorol Zeitschrift* **2006**, *15*, 259–63. <https://doi.org/10.1127/0941-2948/2006/0130>.
 36. Kalombo, K.D. Évolution des éléments du Climat en RDC: Stratégies D'adaptation des Communautés de Base, face aux événements Climatiques de Plus en Plus Fréquents; Éditions Universitaires Européennes: Sarrebruck, Allemagne, **2016**; p. 220.
 37. Munyemba, K.F. ; Bogaert J. Anthropisation et dynamique spatiotemporelle de l'occupation du sol dans la région de Lubumbashi entre 1956 et 2009. *E-Revue UNILU* **2014**, *1*,3–23.
 38. Kasongo, L.M.E.; Baert, B.; Ngongo, L.M.; Verdoodt, A.; Van Ranst, E. Esquisse des pédopaysages du Katanga et aperçu sur les potentialités agricoles et activités minières. In *Anthropisation des Paysages Katangais*; Bogaert, J., Colinet, G., Mahy, G., Eds.; Les Presses Universitaires de Liège—Agronomie: Gembloux, Belgium, **2018**; pp. 127–156.
 39. Dibwe, M.D. Lubumbashi, Ville Industrielle Attractive et Repulsive (1910–2008); Los Restos de la Pobreza Urbana: Madrid, Spain, **2009**.
 40. Useni, S.Y. ; Kaleba, S.C. ; Khonde, C.N. ; Mwana, Y.A. ; Malaisse, F. ; Bogaert, J. & Kankumbi, F.M. Vingt-cinq ans de monitoring de la dynamique spatiale des espaces verts en réponse à (urbanisation dans les communes de la ville de Lubumbashi (Haut-Katanga, RD Congo). *Tropicultura* **2017**, *35*, 300-311.
 41. Main-Knorn, M.; Pflug, B.; Louis, J.; Debaecker, V.; Müller-Wilm, U.; Gascon, F. Sen2Cor for Sentinel-2. *Image Sign. Process. Remote Sens.*, **2017**, *X23*, 10427, 37-48. <https://doi.org/10.1117/12.2278218>.
 42. Misra, G.; Cawkwell, F.; Wingler, A. Status of phenological research using sentinel-2 data: A review. *Remote Sens.* **2020**, *12*, 10–4. <https://doi.org/10.3390/RS12172760>.
 43. Segarra, J. Buchailot, M.L.; Araus, J.L.; Kefauver, S.C. Remote sensing for precision agriculture: Sentinel-2 improved features and applications. *Agronomy* **2020**, *10*, 1–18. <https://doi.org/10.3390/agronomy10050641>.
 44. Wasniewski, A.; Hoscilo, A.; Zagajewski, B.; Moukétou-Tarazewicz, D. Assessment of sentinel-2 satellite images and random forest classifier for rainforest mapping in Gabon. *Forests* **2020**, *11*,1–17. <https://doi.org/10.3390/f11090941>.
 45. Yin, F. ; Lewis, P.E. ; Gómez-Dans, J.L. Bayesian atmospheric correction over land: Sentinel-2/MSI and Landsat 8/OLI. *Geosci Model. Dev.* **2022**, *15*, 7933–76. <https://doi.org/10.5194/gmd-15-7933-2022>.
 46. Barima, Y.S.S. ; Barbier, N. ; Bamba, I. ; Traoré, D. ; Lejoly, J. ; Bogaert, J. Dynamique paysagère en milieu de transition forêt-savane ivoirienne. *Bois For. Trop.* **2009**, *299*, 15. <https://doi.org/10.19182/bft2009.299.a20419>.
 47. Useni, S.Y.; Mpanda, M.M.; Malaisse, F.; Kaseya, P.K.; Bogaert, J. The Spatiotemporal Changing Dynamics of Miombo Deforestation and Illegal Human Activities for Forest Fire in Kundelungu National Park, Democratic Republic of the Congo. *Fire* **2023**, *6*, 174. <https://doi.org/10.3390/fire6050174>.
 48. Useni, S.Y.; Mpanda, M.M.; Khoji, M.H.; Cirezi, C.N.; Malaisse, F.; Bogaert, J. Vegetation Fires in the Lubumbashi Charcoal Production Basin (The Democratic Republic of the Congo): Drivers, Extent and Spatiotemporal Dynamics. *Land* **2023**, *12*, 2171. <https://doi.org/10.3390/land12122171>.
 49. Noi Phan, T.; Kuch, V.; Lehnert, L.W. Land cover classification using google earth engine and random forest classifier—the role of image composition. *Remote Sens.* **2020**, *12*. <https://doi.org/10.3390/RS12152411>.
 50. Belgiu, M.; Drăguț, L. Random forest in remote sensing: A review of applications and future directions. *ISPRS J. Photogramm. Remote Sens.*, **2016**, *114*, 24-31. <https://doi.org/10.1016/j.isprsjprs.2016.01.011>
 51. Gorelick, N.; Hancher, M.; Dixon, M.; Ilyushchenko, S.; Thau, D.; Moore R. Google Earth Engine: Planetary-scale geospatial analysis for everyone. *Remote Sens. Environ.* **2017**, *202*, 18–27. <https://doi.org/10.1016/j.rse.2017.06.031>.

52. Olofsson, P.; Foody, G.M.; Herold, M.; Stehman, S.V.; Woodcock, C.E.; Wulder, M.A. Good practices for estimating area and assessing accuracy of land change. *Remote Sens. Environ.* **2014**, *148*, 42–57. <https://doi.org/10.1016/j.rse.2014.02.015>.
53. Hu, J.; Xia, G.-S.; Hu, F.; Zhang, L. A Comparative Study of Sampling Analysis in the Scene Classification of Optical High-Spatial Resolution Remote Sensing Imagery. *Remote Sens.* **2015**, *7*, 14988–15013. <https://doi.org/10.3390/rs71114988>
54. O'Neill, R.V.; Krummel, J.R.; Garner, R.H.; Sugihara, G.; Jackson, B.; Deangelis, D.L.; Milne, B.T.; Turner, M.G.; Zygmunt, B.; Christensen, S.W.; et al. Indices of landscape pattern. *Landsc. Ecol.* **1988**, *1*, 153–162. <https://doi.org/10.1007/BF00162741>
55. Lamine, S. ; Petropoulos, G.P. ; Singh, S.K. ; Szabó, S. ; Bachari, N.E.I.; Srivastava, P.K.; Suman, S.. Quantifying land use/land cover spatio-temporal landscape pattern dynamics from Hyperion using SVMs classifier and FRAGSTATS®. *Geocarto Int.*, **2018**, *33*(8), 862-878. <https://doi.org/10.1080/10106049.2017.1307460>
56. Cabala, K.S.; Useni, S.Y.; Mwana, Y.A.; Bogaert, J.; Munyemba, K.F. Analyse structurale de la dynamique forestiere dans la région de l'Arc Cuprifère Katangai I. Pattern Analysis of Forest Dynamics of the Katangese Copper Belt Region in the Democratic Republic of the Congo: II. Complementary Analysis on Forest Fragmentation. *Tropicultura* **2018**, *36*, 621–30.
57. Kabuanga, J.M. ; Adipalina Guguya, B. ; Ngenda Okito, E. ; Maestriperi, N. ; Saqalli, M. ; Rossi, V. ; Iyongo Waya Mongo, L. Suiivi de l'anthropisation du paysage dans la région forestière de Babagulu, République Démocratique du Congo. *VertigO*, **2020**, *20*. <https://doi.org/10.4000/vertigo.28347>.
58. Bogaert, J.; Ceulemans, R.; Salvador-Van, E.D. Decision tree algorithm for detection of spatial processes in landscape transformation. *Environ. Manag.* **2004**, *33*, 62–73. <https://doi.org/10.1007/s00267-003-0027-0>
59. de Haulleville, T.; Rakotondraso, O.L.; Rakoto Ratsimba, H.; Bastin, J.F.; Brostaux, Y.; Verheggen, F.J.; Rajoelison, G.L.; Malaisse, F.; Poncelet, M.; Haubruge, É.; Beeckman, H. Fourteen years of anthropization dynamics in the Uapaca bojeri Baill. Forest of Madagascar. *Landsc. Ecol. Eng.* **2018**, *14*, 135–46. <https://doi.org/10.1007/s11355-017-0340-z>.
60. Freitas, S.R.; Hawbaker, T.J.; Metzger, J.P. Effects of roads, topography, and land use on forest cover dynamics in the Brazilian Atlantic Forest. *Forest Ecol. Manag.*, **2010**, *259*(3), 410-417. <https://doi.org/10.1016/j.foreco.2009.10.036>
61. Morgan, J.L.; Gergel, S.E.; Coops, N.C. Aerial photography: A rapidly evolving tool for ecological management. *Bioscience* **2010**, *60*, 47–59. <https://doi.org/10.1525/bio.2010.60.1.9>.
62. Abdi, A.M. Land cover and land use classification performance of machine learning algorithms in a boreal landscape using Sentinel-2 data. *GIScience. Remote Sens.* **2020**, *57*, 1–20. <https://doi.org/10.1080/15481603.2019.1650447>.
63. Wickham, J. Riitters, K.H. Influence of high-resolution data on the assessment of forest fragmentation. *Landsc. Ecol.* **2019**, *34*, 2169–82. <https://doi.org/10.1007/s10980-019-00820-z>.
64. Kim, D.H.; Sexton, J.O.; Noojipady, P.; Huang, C.; Anand, A.; Channan, S.; Feng, M.; Townshend, J.R. Global, Landsat-based forest-cover change from 1990 to 2000. *Remote Sens. Environ.* **2014**, *155*, 178–93. <https://doi.org/10.1016/j.rse.2014.08.017>.
65. López-Serrano, P.M.; Corral-Rivas, J.J.; Díaz-Varela, R.A.; Álvarez-González, J.G.; López-Sánchez, C.A. Evaluation of radiometric and atmospheric correction algorithms for aboveground forest biomass estimation using landsat 5 TM data. *Remote Sens.* **2016**, *8*, 1–19. <https://doi.org/10.3390/rs8050369>.
66. McGarigal, K.; Cushman, S.A.; Neel, M.C.; Ene, E. FRAGSTATS: spatial pattern analysis program for categorical maps. Computer software program produced by the authors at the University of Massachusetts, Amherst **2002**.
67. Zuluaga, S.; Speziale, K.; Lambertucci, S.A. Global Aerial Habitat Conservation Post-COVID-19 Anthro pause. *Trends Ecol. Evol.* **2021**, *36*, 273–7. <https://doi.org/10.1016/j.tree.2021.01.009>.
68. Ferrannini, A.; Barbieri, E.; Biggeri, M.; Di Tommaso, M.R. Industrial policy for sustainable human development in the post-Covid19 era. *World Dev.* **2021**, *137*, 105215. <https://doi.org/10.1016/j.worlddev.2020.105215>.
69. Assefa, N.; Soura, A.; Hemler, E.C.; Korte, M.L.; Wang, D.; Abdullahi, Y.Y. ; Lankoande, B., Millogo, O., Chukwu, A., Workneh, F.; Sie, A. Covid-19 knowledge, perception, preventive measures, stigma, and mental health among healthcare workers in three Sub-Saharan African countries: A phone survey. *Am. J. Trop. Med. Hyg.* **2021**, *105*, 342–50. <https://doi.org/10.4269/ajtmh.20-1621>.
70. Vroh, B.T.A. Plant species used in traditional medicine against the main symptoms of COVID-19 in sub-saharan Africa: Literature review. *Ethnobot. Res. Appl.* **2020**, *20*, 1–14. <https://doi.org/10.32859/era.20.26.1-14>.
71. Duguma, L.A.; van Noordwijk, M.; Minang, P.A.; Muthee, K. COVID-19 pandemic and agroecosystem resilience: Early insights for building better futures. *Sustainability* **2021**, *13*, 1–22. <https://doi.org/10.3390/su13031278>.

72. Mulumbati, M.C.; Godoy Jara, M.; Baboy Longanza, L.; Bogaert, J.; Werbrouck, S.; Sikuzani, Y.U.; Mazinga Kwey, M. In Vitro Regeneration Protocol for *Securidaca longepedunculata* Fresen., a Threatened Medicinal Plant within the Region of Lubumbashi (Democratic Republic of the Congo). *Conservation* **2023**, *3*, 411-425. <https://doi.org/10.3390/conservation3030028>.
73. Lawler, O.K. ; Allan, H.L. ; Baxter, P.W.J. ; Castagnino, R. ; Tor, M.C. ; Dann, L.E.; Hungerford, J.; Karmacharya, D.; Lloyd, T.J.; López-Jara, M.J.; Massie, G.N. The COVID-19 pandemic is intricately linked to biodiversity loss and ecosystem health. *Lancet Planet Heal.* **2021**;5:e840–50. [https://doi.org/10.1016/S2542-5196\(21\)00258-8](https://doi.org/10.1016/S2542-5196(21)00258-8).
74. Chuimika, M.M.; Tshomba, K.J.; Bakari, A.S.; Useni, S.Y.; Werbrouck, S.; Mazinga, MK. Causes de la disparition des plantes médicinales du Miombo Katangais (RD Congo): Cas du commerce non conventionnel de *Securidaca longepedunculata* Fresen (Polygalaceae). *Rev Afr. Environ. Agr.* **2023**, *6*,98–107.
75. Ngamsou Abdel, K. ; Mala, W.A. ; Chimi, P.M. ; Funwi, F.P. ; Engoulou, C. ; Messi Effa, J.A. ; Kamdem, M.E.K. ; Djonko, F.N. ; Fokoua, U.L. ; Adounga, S.B.; Mbolo, M.M. Ethnobotanical study and vulnerability of medicinal plants used against the symptoms of COVID-19 in the Lomié subdivision, East Region of Cameroon. *Heliyon* **2024**, *10*:e28247. <https://doi.org/10.1016/j.heliyon.2024.e28247>.
76. Attah, A.F.; Fagbemi, A.A.; Olubiyi, O.; Dada-Adegbola, H.; Oluwadotun, A.; Elujoba, A.; Babalola, C.P. Therapeutic Potentials of Antiviral Plants Used in Traditional African Medicine With COVID-19 in Focus: A Nigerian Perspective. *Front. Pharmacol.* **2021**, *12*, 1–38. <https://doi.org/10.3389/fphar.2021.596855>.
77. de Sousa, C.; Fatoyinbo, L.; Honzák, M.; Wright, T.M.; Murillo Sandoval, P.J.; Whapoe, Z.E.; Yonmah, J.; Olatunji, E.T.; Garteh, J.; Stovall, A.; Neigh, C.S. Two decades of land cover change and forest fragmentation in Liberia: Consequences for the contribution of nature to people. *Conserv. Sci. Pract.* **2023**, *5*,1–17. <https://doi.org/10.1111/csp2.12933>.
78. Bogaert, J.; Barima, Y.S.S.; Iyongo, W.M.L.; Bamba, I.; Mama, A.; Toy, M.; Lafortezza, R. Forest fragmentation: Causes, ecological impacts, and implications for landscape management. In Li, C., Lafortezza, R., Chen, J., Eds.; Springer: Beijing, China; Berlin/Heidelberg, Germany, **2011**; pp. 273–296.
79. FAO. 2023. Évaluation finale du projet «Gestion communautaire des forêts miombo dans le sud-est du Katanga». Série évaluation de projet, 24/2023 Rome. <https://doi.org/10.4060/cc7156fr>.
80. Ellis, E.A.; Romero Montero, J.A.; Hernández Gómez, I.U. Deforestation processes in the state of quintana roo, mexico: The role of land use and community forestry. *Trop. Conserv. Sci.* **2017**, *10*. <https://doi.org/10.1177/1940082917697259>.
81. Santika, T.; Meijaard, E.; Budiharta, S.; Law, E.A.; Kusworo, A.; Hutabarat, J.A.; Indrawan, T.P.; Struebig, M.; Raharjo, S.; Huda, I. and Ekaputri, A.D. Community forest management in Indonesia: Avoided deforestation in the context of anthropogenic and climate complexities. *Glob. Environ. Chang.* **2017**, *46*, 60–71. <https://doi.org/10.1016/j.gloenvcha.2017.08.002>
82. Pouliot, M.; Treue, T. Rural People’s Reliance on Forests and the Non-Forest Environment in West Africa: Evidence from Ghana and Burkina Faso. *World Dev* **2013**, *43*, 180–93. <https://doi.org/10.1016/j.worlddev.2012.09.010>.
83. Juniyanti, L.; Purnomo, H.; Kartodihardjo, H.; Prasetyo, L.B. Understanding the Driving Forces and Actors of Land Change Due to Forestry and Agricultural Practices in Sumatra and Kalimantan: A Systematic Review. *Land* **2021**, *10*, 463. <https://doi.org/10.3390/land10050463>.
84. Sola, P.; Schure, J.; Eba’a Atyi, R.; Gumbo, D.; Okeyo, I.; Awono, A. Woodfuel policies and practices in selected countries in Sub-Saharan Africa—A critical review. *Bois For. Trop.* **2019**, *340*, 5-19. <https://doi.org/10.19182/bft2019.340.a31690>
85. Kabulu, D.J.-P.; Vranken, I.; Bastin, J.-F.; Malaisse, F.; Nyembwe, N.S.; Useni, S.Y.; Ngongo, L.M.; Bogaert, J. Approvisionnement En Charbon de Bois Des Ménages Lushois: Quantités, Alternatives et Conséquences. In *Anthropisation des Paysages Katangais*; Bogaert, J., Colinet, G., Mahy, G., Eds.; Presses Universitaires de Liège: Liège, Belgique, **2018**; pp. 297–311.
86. Trefon, Th.; Kabuyaya, N. "Les espaces périurbains en Afrique centrale." *Territoires périurbains: Développement, enjeux et perspectives dans les pays du Sud*, Gembloux, les Presses Agronomiques de Gembloux **2015** ; pp 33-42.
87. Useni, S.Y.; Sambieni, K.R.; Maréchal, J.; Ilunga wa Ilunga, E. ; Malaisse, F. ; Bogaert, J. ; Munyemba, K.F. Changes in the Spatial Pattern and Ecological Functionalities of Green Spaces in Lubumbashi (the Democratic Republic of Congo) in Relation With the Degree of Urbanization. *Trop Conserv Sci* **2018**, *11*. <https://doi.org/10.1177/1940082918771325>.
88. Karlson, M.; Mörtberg, U.; Balfors, B. Road ecology in environmental impact assessment. *Environ. Impact Assess. Rev.*, **2014**, *48*, 10-19. <https://doi.org/10.1016/j.eiar.2014.04.002>
89. Branch, A.; Agyei, F.K.; Anai, J.G.; Apecu, S.L.; Bartlett, A.; Brownell, E.; Caravani, M.; Cavanagh, C.J.; Fennell, S.; Langole, S. and Mabele, M.B. From crisis to context: Reviewing the future of sustainable charcoal in Africa. *Energy Res. Soc. Sci.* **2022**, *87*, 102457. <https://doi.org/10.1016/j.erss.2021.102457>.
90. Useni, S.Y.; Khoji, M.H.; Langunu, S.; Gerardy, A.; Bogaert, J. Amplification of Anthropogenic Pressure Heavily Hampers Natural Ecosystems Regeneration within the Savanization Halo Around Lubumbashi

- City (Democratic Republic of Congo). *Int. J. Environ. Sci. Nat. Res.* **2019**, 17(2), 555958. <https://doi.org/10.19080/IJESNR.2019.17.555958>
91. Li, M. ; De Pinto, A. ; Ulimwengu, J. M. ; You, L. ; Robertson, R.D. Impacts of road expansion on deforestation and biological carbon loss in the Democratic Republic of Congo. *Environ. Resou. Econom.* **2015**, 60, 433-469. <https://doi.org/10.1007/s10640-014-9775-y>
 92. Bogaert, J.; Vranken, I.; André, M. Anthropogenic effects in landscapes: Historical context and spatial pattern. In *Biocultural landscapes. Diversity, Functions and Values*; Hong, S.K., Bogaert, J., Min, Q., Eds.; Springer: Dordrecht, The Netherlands, **2014**; pp. 89–112.
 93. Cabala, K.S.; Useni, S.Y.; Sambieni, K.R.; Bogaert, J.; Munyemba, K.F. Dynamique des écosystèmes forestiers de l’Arc Cuprifère Katangais en République Démocratique du Congo. Causes, Transformations spatiales et ampleur. *Tropicultura* **2017**, 35, 192–202.
 94. Haddad, N.M.; Brudvig, L.A. ; Clobert, J. ; Davies, K.F. ; Gonzalez, A. ; Holt, R.D. ; Lovejoy, T.E. ; Sexton, J.O. ; Austin, M.P. ; Collins, C.D.; Cook, W.M. Habitat fragmentation and its lasting impact on Earth’s ecosystems. *Sci. adv.* **2015**, 1, p.e1500052. <https://doi.org/10.1126/sciadv.1500052>.
 95. Laurance, W.F.; Camargo, J.L.; Luizão, R.C.; Laurance, S.G.; Pimm, S.L.; Bruna, E.M.; Stouffer, P.C.; Williamson, G.B.; Benítez-Malvido, J.; Vasconcelos, H.L.; Van Houtan, K.S. The fate of Amazonian forest fragments: a 32-year investigation. *Biol. Conserv.* **2011**, 144, pp.56-67. <https://doi.org/10.1016/j.biocon.2010.09.021>.
 96. Bodo, T.; Gimah, B.G.; Seomoni, K. J. Deforestation and habitat loss: Human causes, consequences and possible solutions. *J. Geogr. Res.* **2021**, 4, 22-30. <https://doi.org/10.30564/jgr.v4i2.3059>.
 97. Fischer, R.; Taubert, F.; Müller, M. S.; Groeneveld, J.; Lehmann, S.; Wiegand, T. & Huth, A. Accelerated forest fragmentation leads to critical increase in tropical forest edge area. *Sci. adv.* **2021**, 7, eabg7012. <https://doi.org/10.1126/sciadv.abg7012>
 98. McNicol, I. M.; Keane, A.; Burgess, N. D.; Bowers, S. J.; Mitchard, E.T.; Ryan, C.M. Protected areas reduce deforestation and degradation and enhance woody growth across African woodlands. *Commun. Earth Environ.* **2023**, 4, 392. <https://doi.org/10.1038/s43247-023-01053-4>.
 99. Dieng, M.; Mbow, C.; Skole, D.L.; Ba, B. Sustainable land management policy to address land degradation: linking old forest management practices in Senegal with new REDD+ requirements. *Front. environ. sci.* **2023**, 11, 1088726. <https://doi.org/10.3389/fenvs.2023.1088726>.
 100. Mbuvi, M.T.E.; Kungu, J.B. A transforming traditional community based forest management: the case of Loita community forest, Kenya. *Heliyon* **2021**, 7. <https://doi.org/10.1016/j.heliyon.2021.e07380>.
 101. Abiodun, B. J.; Adeyewa, Z. D.; Oguntunde, P. G.; Salami, A. T.; Ajayi, V. O. Modeling the impacts of reforestation on future climate in West Africa. *Theor. Appl. Climatol.* **2012**, 110, 77-96. <https://doi.org/10.1007/s00704-012-0614-1>
 102. Negi, S.; Pham, T. T.; Karky, B.; Garcia, C. Role of community and user attributes in collective action: case study of community-based forest management in Nepal. *Forests* **2018**, 9, 136. <https://doi.org/10.3390/f9030136>
 103. Panja, P. Deforestation, Carbon dioxide increase in the atmosphere and global warming: A modelling study. *Int. J. Model. Simul.* **2021**, 41, 209-219. <https://doi.org/10.1080/02286203.2019.1707501>
 104. Tole, L. Reforms from the ground up: a review of community-based forest management in tropical developing countries. *Environ. Manag.*, **2010**, 45(6), 1312-1331. <https://doi.org/10.1007/s00267-010-9489-z>
 105. Li, S.; Xu, M.; Sun, B. Long-term hydrological response to reforestation in a large watershed in southeastern China. *Hydrol. Process.* **2014**, 28, 5573-5582. <https://doi.org/10.1002/hyp.10018>
 106. Peng, H.; Tague, C.; Jia, Y. Evaluating the eco-hydrologic impacts of reforestation in the Loess Plateau, China, using an eco-hydrologic model. *Ecohydrology* **2016**, 9, 498-513. <https://doi.org/10.1002/eco.1652>
 107. Hu, X.; Næss, J. S.; Jordan, C. M.; Huang, B.; Zhao, W.; Cherubini, F. Recent global land cover dynamics and implications for soil erosion and carbon losses from deforestation. *Anthropocene* **2021**, 34, 100291. <https://doi.org/10.1016/j.ancene.2021.100291>
 108. Miyamoto, M. Poverty reduction saves forests sustainably: Lessons for deforestation policies. *World Dev.* **2020**, 127, 104746. <https://doi.org/10.1016/j.worlddev.2019.104746>
 109. Arima, E. Y.; Barreto, P.; Araújo, E.; Soares-Filho, B. Public policies can reduce tropical deforestation: Lessons and challenges from Brazil. *Land use policy* **2014**, 41, 465-473. <https://doi.org/10.1016/j.landusepol.2014.06.026>
 110. Musyoki, J. K.; Mugwe, J.; Mutundu, K.; Muchiri, M. Factors influencing level of participation of community forest associations in management forests in Kenya. *J. Sustain. For.* **2016**, 35, 205-216. <https://doi.org/10.1080/10549811.2016.1142454>
 111. Faria, D.; Morante-Filho, J.C.; Baumgarten, J.; Bovendorp, R.S.; Cazetta, E.; Gaiotto, F.A.; Mariano-Neto, E.; Mielke, M.S.; Pessoa, M.S.; Rocha-Santos, L.; Santos, A.S. The breakdown of ecosystem functionality driven by deforestation in a global biodiversity hotspot. *Biol. Conserv.* **2023**, 283, 110126. <https://doi.org/10.1016/j.biocon.2023.110126>.

112. Mpanda, M.M. ; Kipili, M.I. ; Tshomba, K.J. ; Kitabala, M.A. ; Cabala, K.S. ; Useni, S.Y. Perception de la dégradation de la fertilité des sols et de sa gestion par les agriculteurs de la cité de Kasenga en République Démocratique du Congo. *Geo-Eco-Trop*, **2021**, 45(2), 211-220.
113. Loro, M.; Ortega, E.; Arce, R.M.; Geneletti, D. Ecological connectivity analysis to reduce the barrier effect of roads. An innovative graph-theory approach to define wildlife corridors with multiple paths and without bottlenecks. *Landscape Urban Plan.* **2015**, 139, 149-162. <https://doi.org/10.1016/j.landurbplan.2015.03.006>
114. Kuglerová, L.; Ågren, A.; Jansson, R. & Laudon, H. towards optimizing riparian buffer zones: Ecological and biogeochemical implications for forest management. *For. Ecol. Manag.* **2014**, 334, 74-84. <https://doi.org/10.1016/j.foreco.2014.08.033>.
115. Young, N. E.; Evangelista, P. H.; Mengitsu, T. & Leisz, S. Twenty-three years of forest cover change in protected areas under different governance strategies: A case study from Ethiopia's southern highlands. *Land Use pol.* **2020**, 91, 104426. <https://doi.org/10.1016/j.landusepol.2019.104426>.
116. Tollefson, J. Why deforestation and extinctions make pandemics more likely. *Nature*, **2020**, 584(7820), 175-177. <https://doi.org/10.1038/d41586-020-02341-1>
117. Gebara, M.F.; May, P.H.; Platais, G. Pandemics, conservation, and human-nature relations. *Climate Change Ecol.*, **2021**, 2, 100029. <https://doi.org/10.1016/j.ecochg.2021.100029>
118. Pereira, M.; Oliveira, A. M. Poverty and food insecurity may increase as the threat of COVID-19 spreads. *Public Health Nutr.*, **2020**, 3(17), 3236-3240. <https://doi.org/10.1017/S1368980020003493>.
119. Tajudeen, Y.A.; Oladunjoye, I.O.; Bajinka, O.; Oladipo, H.J. Zoonotic Spillover in an Era of Rapid Deforestation of Tropical Areas and Unprecedented Wildlife Trafficking: Into the Wild. *Challenges* **2022**, 13, 41. <https://doi.org/10.3390/challe13020041>
120. Salomon, W. ; Useni, S.Y. ; Kouakou, A.T.M. ; Barima, Y.S.S. ; Theodat, J.M. ; Bogaert, J. Cartographie et quantification de la perte du couvert forestier dans les parcs nationaux du Sud de la République d'Haïti de 1985 à 2018: rôle des activités agricoles locales. *Int. J. Biol. Chem. Sci.*, **2021**, 15(2), 477-496. <https://dx.doi.org/10.4314/ijbcs.v15i2.9>
121. Berrouet, L. M.; Machado, J.; Villegas-Palacio, C. Vulnerability of socio—ecological systems: A conceptual Framework. *Ecol. Indic.*, **2018**, 84, 632-647. <https://doi.org/10.1016/j.ecolind.2017.07.051>

Disclaimer/Publisher's Note: The statements, opinions and data contained in all publications are solely those of the individual author(s) and contributor(s) and not of MDPI and/or the editor(s). MDPI and/or the editor(s) disclaim responsibility for any injury to people or property resulting from any ideas, methods, instructions or products referred to in the content.